

**Lower Lakes Aquatic and Littoral Vegetation
Condition Monitoring
2009-10**



Susan Gehrig, Jason Nicol and Kelly Marsland

**SARDI Publication No. F2010/000370-2
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Management Board



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Executive Summary

Condition Monitoring Plan for the Coorong, Lower Lakes and Murray Mouth Icon Site (Maunsell Australia Pty Ltd. 2009) identifies that a monitoring program is required that focuses on measuring the key environmental parameters that are indicators of river and floodplain health in both the short and longer terms. This report presents the findings of the first two years of a monitoring program established to evaluate V3 in the aforementioned plan: maintain or improve aquatic and littoral vegetation in the Lower Lakes.

Vegetation surveys were conducted at selected wetlands and lakeshore sites across Lakes Alexandrina and Albert, Goolwa Channel, the lower Finniss River and lower Currency Creek and the mouths of the Angas and Bremer Rivers. Sites established in spring 2008 were resurveyed and additional sites were set up in spring 2009 in Goolwa Channel, the lower Finniss River and lower Currency Creek to assess the impact of the Clayton regulator. At each site, transects were established perpendicular to the shoreline and three 1 x 3 m quadrats separated by 1 m were located at regular elevation intervals (defined by plant community) for wetlands or elevations (+0.8, +0.6, +0.4, +0.2, 0 and -0.5 m AHD) for lakeshores. The cover and abundance of each species present in quadrats was estimated using a modified Braun-Blanquet (1932) cover abundance score. Vegetation surveys were undertaken in October 2008, March 2009, October 2009 and March 2010 during a period of record low water levels in the Lower Lakes. However, significant engineering interventions (construction of the Clayton regulator and pumping of Narrung Wetland) also occurred during the study period and were assessed as part of the monitoring program.

Between spring 2008 and autumn 2010, a total of 127 taxa (including 67 exotics) were recorded. Disconnection and subsequent desiccation of wetlands generally resulted in the loss of submergent taxa (except from wetlands that received rainfall runoff or were filled by pumping i.e. Narrung) and colonisation of terrestrial species. In 2009-10 more terrestrial species colonised lower elevations than in 2008-09. Low water levels resulted in colonisation of terrestrial species (lower elevations were colonised in 2009-10 compared with 2008-09) around the shorelines of Lakes Alexandrina and Albert, similar to wetlands. The majority of the terrestrial taxa that colonised wetlands and lakeshores were exotic agricultural weeds (generally pasture grasses and legumes), which is probably due to surrounding land use. In contrast, a diverse amphibious, emergent and submergent plant community had developed by March 2010 in areas inundated by the Clayton regulator, which in 2008-09 were dominated by terrestrial taxa and bare soil.

Comparison between monitoring results and the River Murray Wetlands baseline surveys undertaken in 2004 and 2005 show that target V3 has not been met for understorey vegetation, except in Goolwa Channel, the lower Finniss River, lower Currency Creek and Narrung

Wetland. Results from Narrung Wetland and areas inundated by the Clayton regulator provide insight into the recovery of the system when water levels are restored to historical levels and show that the system is resilient and can recover from the current degraded state.

1 Introduction

The Coorong, Lower Lakes and Murray Mouth region has been listed as one of six icon sites under Murray-Darling Basin Authority's "The Living Murray" (ILM) program. The Condition Monitoring Plan for the Coorong, Lower Lakes and Murray Mouth Icon Site outlined a series of 17 condition targets for the Icon Site (Maunsell Australia Pty Ltd. 2009). This report presents the findings from the first two years of the understorey component of a condition monitoring program designed to evaluate target V3: maintain or improve aquatic and littoral vegetation in the Lower Lakes (Marsland and Nicol 2009).

Scientifically defensible and statistically robust monitoring programs need to be established to assist in meeting the ecological targets in the Coorong, Lower Lakes and Murray Mouth Icon Site and the Ramsar Management Plan. Marsland and Nicol (2006) identified that existing monitoring programs (in 2006) would not adequately assess target V3; therefore, a monitoring program that expanded upon existing monitoring programs (SAMDBNRM Board community wetland monitoring) and builds upon this program was established in 2008 (Marsland and Nicol 2009). The understorey vegetation monitoring program described in this report uses the same methods and sites as the SAMDBNRM Board community wetland monitoring program but includes additional sites in lakeshore habitats, the lower reaches of the Finnis River, Currency Creek and Goolwa Channel and wetlands that were not part of the original community wetland monitoring program (Marsland and Nicol 2009).

The Condition Monitoring Plan for the Icon Site proposed 'indicators for monitoring' that comprised of individual taxa and discrete communities: *Melaleuca balmaturorum*, *Myriophyllum* spp., *Gabnia filum*, *Schoenoplectus* spp., *Typha domingensis*, *Phragmites australis* and samphire communities (Maunsell Australia Pty Ltd. 2009). However, further discussions concluded that the entire understorey vegetation would be monitored with a separate technique used for *M. balmaturorum*. Hence, the monitoring program consists of two complementary components: the first component involves the monitoring of aquatic and littoral understorey vegetation in spring (high lake levels) and autumn (low lake levels) to determine the current condition and seasonal changes in floristic composition and the second component monitors the mid to long-term population dynamics of the dominant tree species *Melaleuca balmaturorum*. The *Melaleuca balmaturorum* component of the monitoring program is undertaken every three to five years and stand condition was not monitored in 2009-10. Information regarding *Melaleuca balmaturorum* stand condition is presented in Marsland and Nicol (2009).

Aquatic and littoral plants play important roles in the ecosystem dynamics of the Lower Lakes. Plants are major primary producers (e.g. Noges *et al.* 2010), provide habitat for fish (e.g. Wedderburn *et al.* 2007), birds (e.g. Mondon *et al.* 2009) and invertebrates (e.g. Cronin *et al.*

2006), stabilise sediments (e.g. Abernethy and Rutherford 1998) and improve water quality (Rodriguez-Gallego *et al.* 2010). Historically (post barrage construction (1940) to 2007) the Lower Lakes fringing wetlands were permanently inundated and contained diverse submergent, emergent and amphibious plant communities (Renfrey *et al.* 1989; Holt *et al.* 2005; Nicol *et al.* 2006). Lakeshore habitats were dominated by large stands of emergent macrophytes that had recruited along the shoreline with submergent and amphibious species restricted to sheltered areas with low wave intensity (e.g. Milang, Goolwa Channel, the lower Finniss River and lower Currency Creek) (J. Nicol pers. obs.).

Water levels in Lakes Alexandrina and Albert have fallen to unprecedented lows (<-0.75 m AHD) since 2007, disconnecting fringing wetlands and exposing extensive areas of acid sulfate soils (Merry *et al.* 2003; Fitzpatrick *et al.* 2009a; Fitzpatrick *et al.* 2009b). The threat of acid sulfate soil development due to low water levels prompted two large-scale engineering interventions: the construction of a bund at Narrung (completed in May 2008) and a regulator at Clayton (completed in August 2009). The bund at Narrung was constructed to prevent water levels in Lake Albert falling below -0.5 m AHD (water was pumped from Lake Alexandrina to maintain Lake Albert water levels). The regulator at Clayton was constructed to impound flows from the Finniss River and Currency Creek between the regulator and Goolwa Barrage; in addition water was pumped from Lake Alexandrina to raise water levels to +0.7 m AHD in Goolwa Channel. Construction of the Narrung and Clayton regulators has hydrologically disconnected Lake Albert and Goolwa Channel (and the lower Currency Creek and Finniss River) from Lake Alexandrina allowing these water bodies to be operated independently. Therefore, differences in water levels now exist between Lake Alexandrina, Lake Albert and Goolwa Channel in a system where these waterbodies were historically connected. This may have particular consequences for plant communities (e.g. Rea and Ganf 1994; Brock and Casanova 1997; Nicol *et al.* 2003). An additional hydrological intervention was undertaken at Narrung Wetland, 250 ML of environmental water from Lake Alexandrina was pumped into the wetland in October 2009 to provide suitable conditions for the growth of submergent taxa (particularly *Ruppia tuberosa*).

Results from 2008-09 showed that disconnection and subsequent desiccation of fringing wetlands changed the plant community from a diverse amphibious, emergent and submergent community (Renfrey *et al.* 1989; Holt *et al.* 2005; Nicol *et al.* 2006) to one dominated by terrestrial taxa or bare soil (Marsland and Nicol 2009). Terrestrial species also dominated lakeshore habitats in Lake Alexandrina and Goolwa Channel; however, large stands beds of emergent species were still present along the historical shoreline and the lowest elevation surveyed (-0.5 m AHD) was generally bare (Marsland and Nicol 2009). Despite the stable water levels, results were similar for Lake Albert with no recruitment of shoreline vegetation at -0.5 m AHD (Marsland and Nicol 2009).

The monitoring undertaken in 2009-10 builds on data collected in 2008-09 and provides information regarding the change in plant communities since spring 2008. The study period includes a period of record low water levels in Lake Alexandrina but also includes an engineering intervention that has resulted in elevated water levels in Goolwa Channel. Therefore, this monitoring program will simultaneously collect information regarding the change in wetland plant communities in response to drawdown, desiccation and increasing water levels. Consequently it will provide an insight into recovery of the system when water levels are restored to historical levels. The aims of this project were to:

- continue the statistically robust, quantitative understory aquatic and littoral vegetation monitoring program in the Lower Lakes to assess TLM target V3
- expand monitoring program (in conjunction with a complementary project funded by DENR) to investigate the impact of increased water levels in Goolwa Channel, the lower Finniss River and lower Currency Creek

2 Methods

Monitoring of understory vegetation was conducted at 11 wetland and 25 lakeshore sites in October 2008, March 2009, October 2009 and March 2010 (for sites established in 2008 or earlier) and October 2009 and March 2010 for sites established in 2009 (Table 1, Figure 1). Sites were grouped on the basis of habitat (lakeshore or wetland) and location (Lake Alexandrina, Lake Albert or Goolwa Channel). GPS coordinates for each site are listed in Appendix 1.

Table 1: List of understory vegetation site numbers (relative to map provided in Figure 1), site name, location, habitat type (wetland or lakeshore) and the year sites were established.

| Site # | Site Name | Location | Habitat | Year Established |
|--------|-------------------------------|------------------|-----------|------------------|
| 1 | Bremer Mouth Lakeshore | Lake Alexandrina | lakeshore | 2008 |
| 2 | Brown Beach 1 | Lake Albert | lakeshore | 2008 |
| 3 | Brown Beach 2 | Lake Albert | lakeshore | 2008 |
| 4 | Clayton Bay | Goolwa Channel | lakeshore | 2009 |
| 5 | Currency Creek 3 | Goolwa Channel | lakeshore | 2008 |
| 6 | Currency Creek 4 | Goolwa Channel | lakeshore | 2008 |
| 7 | Goolwa North | Goolwa Channel | lakeshore | 2009 |
| 8 | Goolwa South | Goolwa Channel | lakeshore | 2009 |
| 9 | Hindmarsh Island Bridge 01 | Goolwa Channel | lakeshore | 2009 |
| 10 | Hindmarsh Island Bridge 02 | Goolwa Channel | lakeshore | 2009 |
| 11 | Lake Reserve Rd | Lake Alexandrina | lakeshore | 2008 |
| 12 | Loveday Bay | Lake Alexandrina | lakeshore | 2009 |
| 13 | Loveday Bay Lakeshore | Lake Alexandrina | lakeshore | 2009 |
| 14 | Lower Finnis 02 | Goolwa Channel | lakeshore | 2009 |
| 15 | Milang | Lake Alexandrina | wetland | pre-2008 |
| 16 | Milang Lakeshore | Lake Alexandrina | lakeshore | 2009 |
| 17 | Pt Sturt Lakeshore | Lake Alexandrina | lakeshore | 2008 |
| 18 | Pt Sturt Water Reserve | Lake Alexandrina | lakeshore | 2008 |
| 19 | Teringie Lakeshore | Lake Alexandrina | lakeshore | 2008 |
| 20 | Upstream of Clayton Regulator | Lake Alexandrina | lakeshore | 2009 |
| 21 | Wally's Landing | Goolwa Channel | lakeshore | 2009 |
| 22 | Warrengie 1 | Lake Albert | lakeshore | 2009 |
| 23 | Lower Finnis 03 | Goolwa Channel | lakeshore | 2009 |
| 24 | Narrung Lakeshore | Lake Alexandrina | lakeshore | 2008 |
| 25 | Nurra Nurra | Lake Albert | lakeshore | 2008 |
| 26 | Warrengie 2 | Lake Albert | lakeshore | 2009 |
| 27 | Angas Mouth | Lake Alexandrina | wetland | 2008 |
| 28 | Bremer Mouth | Lake Alexandrina | wetland | 2008 |
| 29 | Dunns Lagoon | Lake Alexandrina | wetland | 2008 |
| 30 | Goolwa Channel Drive | Lake Alexandrina | wetland | 2008 |
| 31 | Hunters Creek | Lake Alexandrina | wetland | 2008 |
| 32 | Poltalloch | Lake Alexandrina | wetland | 2008 |
| 33 | Pt Sturt | Lake Alexandrina | wetland | 2008 |
| 34 | Teringie | Lake Alexandrina | wetland | pre-2008 |
| 35 | Waltowa | Lake Albert | wetland | pre-2008 |
| 36 | Narrung | Lake Alexandrina | wetland | pre-2008 |

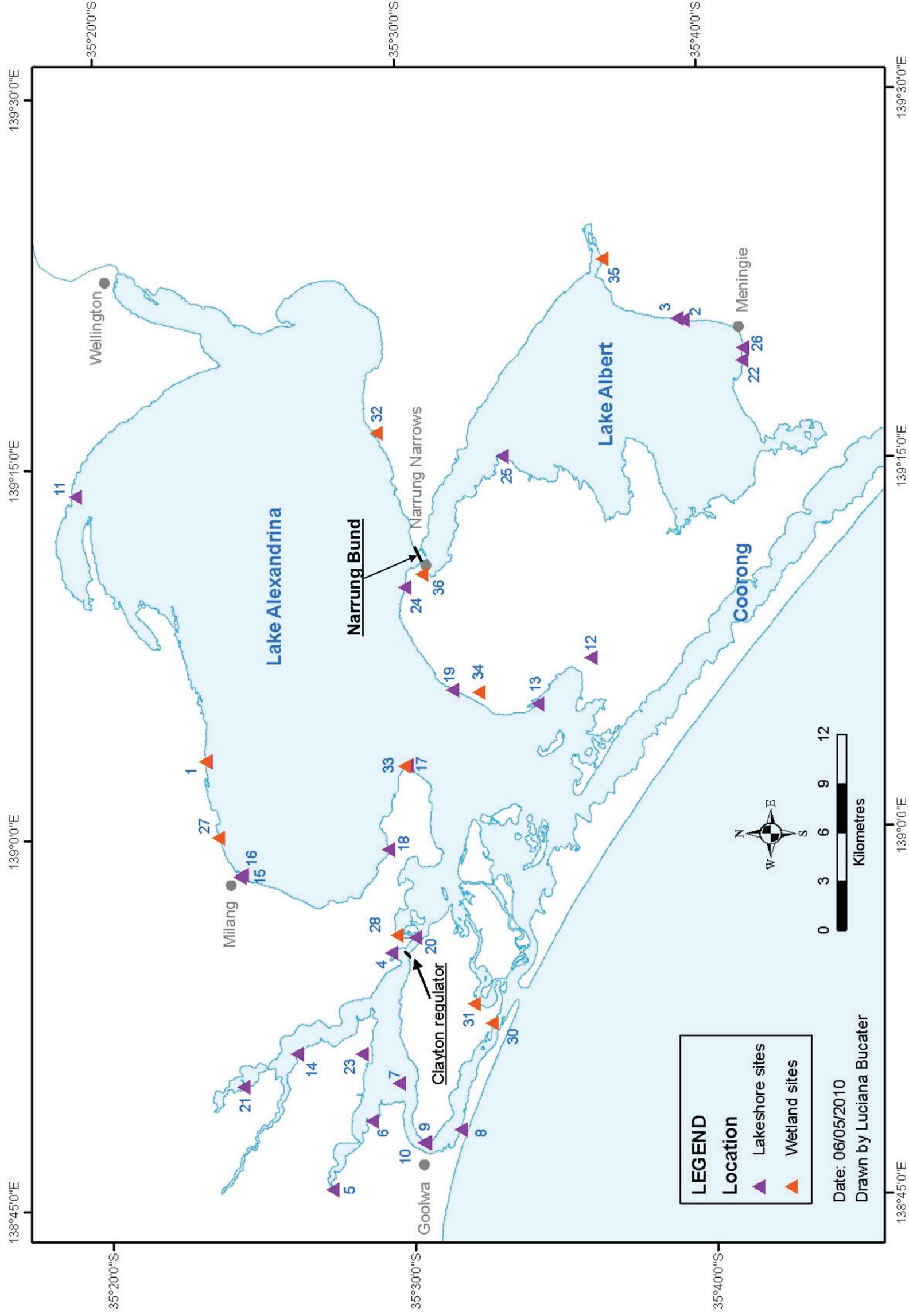


Figure 1: Map of Lakes Alexandrina and Albert and Goolwa Channel showing the location of lakeshore and wetland monitoring sites.

2.1 Vegetation surveying protocol

2.1.1 Wetlands

At each site a transect running perpendicular to the shoreline was established and three 1 x 3 m quadrats separated by 1 m were positioned at regular elevation intervals (A. Frears pers. comm.). In wetlands with an existing monitoring program (Milang, Waltowa, Teringie and Narrung) existing sites were re-surveyed. For the remaining wetlands (Dunns Lagoon, Pt Sturt, Hunters Creek, Goolwa Channel Drive, Bremer River Mouth, Angas River Mouth and Loveday Bay) a transect was established and quadrats placed in each vegetation community present during the spring 2008 survey. A minimum of one additional transect (but usually three or more transects were set up in each wetland) was established and quadrats were placed at the same elevations as on the first transect.

Cover and abundance of each species present in the quadrat were estimated using the method outlined in Heard and Channon (1997) except that N (not many, 1-10 individuals) and T (sparsely or very sparsely present) were replaced by 0.1 and 0.5 to enable statistical analyses (Table 2).

Table 2: Modified Braun-Blanquet (1932) scale estimating cover/abundance as per Heard and Channon (1997).

| Score | Modified Score | Description |
|-------|----------------|--|
| N | 0.1 | Not many, 1-10 individuals |
| T | 0.5 | Sparsely or very sparsely present; cover very small (less than 5%) |
| 1 | 1 | Plentiful but of small cover (less than 5%) |
| 2 | 2 | Any number of individuals covering 5-25% of the area |
| 3 | 3 | Any number of individuals covering 25-50% of the area |
| 4 | 4 | Any number of individuals covering 50-75% of the area |
| 5 | 5 | Covering more than 75% of the area |

2.1.2 Lakeshores

The vegetation surveying protocol used to survey wetlands (see section 2.1.1, Table 2), was also used for lakeshores. However, quadrats were positioned at +0.8, +0.6, +0.4, +0.2, 0 and -0.5 m AHD on transects running perpendicular to the shore.

2.2 Plant identification and nomenclature

Plants were identified using keys in Cunningham *et al.* (1981), Jessop and Toelken (1986), Dashorst and Jessop (1998), Romanowski (1998), Sainty and Jacobs (1981; 2003), Prescott

(1998) and Jessop *et al.* (2006). In some cases due to immature individuals or lack of floral structures plants were identified to genus only. Nomenclature used follows Barker *et al.* (2005).

2.3 Functional Groups

Due to the large number of taxa present, species were classified into functional groups (based on water regime preferences) outlined in Table 3. Functional classification was based on the classification framework devised by Brock and Casanova (1997), which was based on species from wetlands in the New England Tablelands region of New South Wales and modified by Gehrig and Nicol (2010) to suit the Lower Lakes. The position each group occupies in relation to flooding depth and duration is outlined in Figure 2.

The use of a functional group approach to assess change through time and potential impacts of management strategies has several advantages compared to a species or community based approach:

- species with similar water regimes preferences are grouped together, which simplifies systems with high species richness (especially where there are large numbers of species with similar water regime preferences),
- predictions about the response of the plant community are made based on processes; therefore, does not require prior biological knowledge of the system,
- is transferrable between systems,
- robust and testable models that predict the response of a system to an intervention or natural event can be constructed, which can in turn be used as hypotheses for monitoring programs.

However there are limitations of the approach, which include:

- loss of information on species or communities (especially if there are species or communities of conservation significance or there is a pest plant problem),
- uncertainty regarding which species should be classified into which functional group,
- important factors (e.g. salinity) are often not taken into consideration (additional factors can be included. However, this can often complicate the functional classification and in systems where there is low species richness the number of groups may be greater than the number of species).

In this report, changes through time between locations and elevations and TLM targets will be assessed using species and functional approaches.

Table 3: Functional classification of plant species based on water regime preferences, modified from Brock and Casanova (1997).

| Functional Group | Water Regime Preference | Examples |
|--|---|--|
| Terrestrial dry | Will not tolerate inundation and tolerates low soil moisture for extended periods. | <i>Rhagodia spinescens</i> , <i>Enchylaena tomentosa</i> |
| Terrestrial damp | Will tolerate inundation for short periods (<2 weeks) but require high soil moisture throughout their life cycle. | <i>Centaurea calcitrapa</i> , <i>Chenopodium album</i> , <i>Fumaria bastardii</i> |
| Floodplain | Temporary inundation, plants germinate on newly exposed soil after flooding but not in response to rainfall. | <i>Lachnagrostis filiformis</i> |
| Amphibious fluctuation tolerator-emergent | Fluctuating water levels, plants do not respond morphologically to flooding and drying and will tolerate short-term complete submergence (<2 weeks). | <i>Cyperus gymnocaulos</i> , <i>Juncus kraussii</i> , <i>Schoenoplectus pungens</i> |
| Amphibious fluctuation tolerator-woody | Fluctuating water levels, plants do not respond morphologically to flooding and drying and are large perennial woody species. | <i>Melaleuca halmaturorum</i> , <i>Muehlenbeckia florulenta</i> |
| Amphibious fluctuation tolerator-low growing | Fluctuating water levels, plants do not respond morphologically to flooding and drying and are generally small herbaceous species. | <i>Limosella australis</i> , <i>Crassula helmsii</i> , <i>Brachycome basaltica</i> |
| Amphibious fluctuation responder-plastic | Fluctuating water levels, plants respond morphologically to flooding and drying (e.g. increasing above to below ground biomass ratios when flooded). | <i>Persicaria lapathifolium</i> |
| Floating | Static or fluctuating water levels, responds to fluctuating water levels by having some or all organs floating on the water surface. Most species require permanent water to survive. | <i>Azolla</i> spp., <i>Lemna</i> spp., |
| Submergent r-selected | Temporary wetlands that hold water for longer than 4 months. | <i>Ruppia tuberosa</i> , <i>Ruppia polycarpa</i> , <i>Lamprothamnium papulosum</i> |
| Emergent | Static shallow water <1 m or permanently saturated soil. | <i>Typha</i> spp., <i>Phragmites australis</i> , <i>Schoenoplectus validus</i> |
| Submergent k-selected | Permanent water. | <i>Myriophyllum</i> spp. <i>Vallisneria australis</i> , <i>Ruppia megacarpa</i> <i>Potamogeton pectinatus</i> |

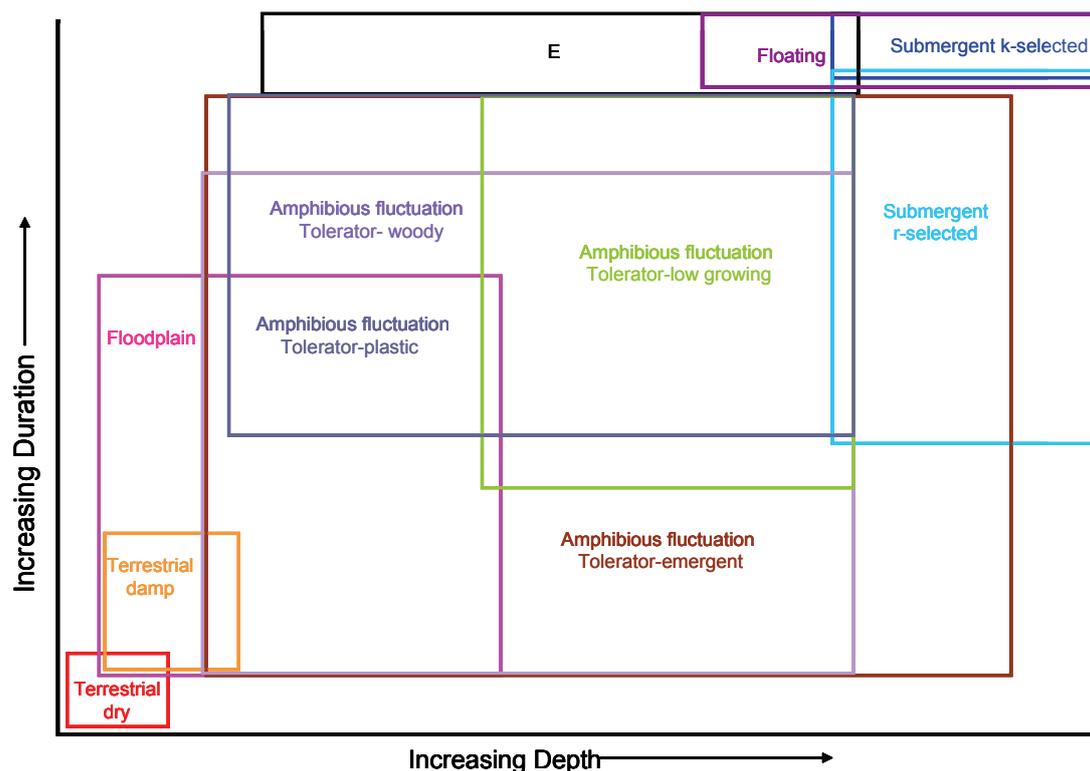


Figure 2: Plant functional groups in relation to depth and duration of flooding.

The “terrestrial dry” functional group is intolerant of flooding and taxa will persist in environments with low soil moisture (Table 3) (Brock and Casanova 1997). Taxa from this functional group often invade wetlands that have been drawn down for an extended period or floodplains where there has been a lack of flooding but are generally restricted to highlands that never flood (Brock and Casanova 1997).

Taxa in the “terrestrial damp” group will tolerate inundation for short periods and require high soil moisture to complete their life cycle (Table 3) (Brock and Casanova 1997). Taxa from this functional group are often winter annuals, perennial species that grow around the edges of permanent water bodies where there is high soil moisture or species that colonise wetlands shortly after they are drawn down and riparian zones and floodplains shortly after flood waters recede (Brock and Casanova 1997).

Taxa in the “floodplain” functional group exhibit most of the traits of terrestrial species; they are generally intolerant of long-term inundation but are restricted to areas that flood periodically (they are absent from the highlands) because they only germinate after flood waters recede or wetlands are drawn down, not in response to rainfall (Table 3) (Nicol 2004). Taxa from this functional group colonise floodplains and riparian zones after flood waters have receded and when wetlands are drawn down (Nicol 2004). Floodplain species often have flexible life history

strategies, they grow whilst soil moisture is high and flower and set seed (after which most species die) in response to low soil moisture (Nicol 2004).

The “amphibious fluctuation tolerator-emergent” group consists mainly of emergent sedges and rushes that prefer high soil moisture or shallow water but require their photosynthetic parts to be emergent, although many will often tolerate short-term submergence (Table 3) (Brock and Casanova 1997). Taxa from this group are often found on the edges of permanent water bodies, in seasonal and temporary wetlands, in riparian zones and areas that frequently wet and dry.

Species in the “amphibious fluctuation tolerator-woody” group have similar water regime preferences to the amphibious fluctuation tolerator-emergent group (Figure 2) and consist of woody perennial species (Table 3) (Brock and Casanova 1997). These plants generally require high soil moisture in the root zone but there are several species that are tolerant of desiccation for extended periods (Roberts and Marston 2000). Species in this functional group are generally found on the edges of permanent water bodies, in seasonal and temporary wetlands, in riparian zones and areas that frequently wet and dry.

The “amphibious fluctuation tolerator-low growing” group have similar water regime preferences to the amphibious fluctuation tolerator-emergent and amphibious fluctuation tolerator-woody group (Figure 2); however, some species can grow totally submerged except during flowering (when there is a requirement for a dry phase) (Table 3) (Brock and Casanova 1997). Species in this functional group are generally found on the edges of permanent water bodies, in seasonal and temporary wetlands, in riparian zones and areas that frequently wet and dry but species are usually less desiccation tolerant than species in the other amphibious tolerator groups (Figure 2).

The “amphibious fluctuation responder-plastic” group occupies a similar zone to the amphibious fluctuation tolerator-low growing group; except that they have a physical response to water level changes such as rapid shoot elongation or a change in leaf type (Brock and Casanova 1997). They can persist on damp and drying ground because of their morphological flexibility but can flower even if the site does not dry out. They occupy a slightly deeper/wet for longer area than the amphibious fluctuation tolerator-low growing group (Figure 2).

Species in the “floating” functional group float on the top of the water (often unattached to the sediment) with the majority of species requiring the presence of free water of some depth year round; although, some species can survive and complete their life cycle stranded on mud (Table 3) (Brock and Casanova 1997). Taxa in this group are usually found in permanent waterbodies, often forming large floating mats upstream of barriers (e.g. weirs), in lentic water bodies and slackwaters.

“Submergent r-selected” species colonise recently flooded areas (Table 3) and show many of the attributes of Grime’s (1979) r-selected (ruderal) species, which are adapted to periodic disturbances. Many require drying to stimulate germination; they frequently complete their life cycle quickly and die off naturally. They persist via a dormant, long-lived bank of seeds, spores or asexual propagules (e.g. *Ruppia tuberosa* and *Ruppia polycarpa* turions in the sediment) (Brock 1982). They prefer habitats that are annually flooded to a depth of more than 10cm but can persist as dormant propagules for a number of years (temporary or ephemeral wetlands).

The “emergent” group consists of taxa that require permanent shallow water or a permanently saturated root zone, but require emergent leaves or stems (Table 3). They are often found on the edges of permanent waterbodies and in permanent water up to 2 m deep (depending on species) or in areas where there are shallow water tables (Roberts and Marston 2000).

“Submergent k-selected” species require permanent water greater than 10 cm deep for more than a year to either germinate or reach sufficient biomass to start reproducing (Table 3) (Roberts and Marston 2000). Species in this group show many of the attributes of Grime’s (1979) k-selected (competitor) species that are adapted to stable environments and are only found in permanent water bodies. The depth of colonisation of submergent k-selected species is dependant on photosynthetic efficiency and water clarity (*sensu* Spence 1982).

2.4 Data Analysis

2.4.1 Wetlands

The changes in floristic composition through time and between elevations were analysed for each wetland (except Angas and Bremer Mouths, which were pooled) using two-factor PERMANOVA (Anderson 2001; Anderson and Ter Braak 2003), NMS ordination, Group Average Clustering (McCune *et al.* 2002) and Indicator Species Analysis (Dufrene and Legendre 1997) using the packages PCOrd version 5.12 (McCune and Mefford 2006) and PRIMER version 6.1.12 (Clarke and Gorley 2006). For Hunter’s Creek, the middle of the wetland was devoid of vegetation, so changes through time were only compared at the edge elevation using single-factor PERMANOVA, (Anderson 2001; Anderson and Ter Braak 2003) and Indicator Species Analysis (Dufrene and Legendre 1997).

2.4.2 Lakeshores

Lakeshore sites were analysed separately to the wetlands. The changes in floristic composition through time and between elevations for each location (Goolwa Channel, Lake Alexandrina and Lake Albert at sites where two year’s of data were available) were analysed using two-factor PERMANOVA (Anderson 2001; Anderson and Ter Braak 2003) and Indicator Species Analysis

(Dufrene and Legendre 1997) using the packages PCOrd version 5.12 (McCune and Mefford 2006) and PRIMER version 6.1.12 (Clarke and Gorley 2006).

Changes in the plant community from spring 2009 and autumn 2010 for each elevation (+0.8, +0.6, +0.4, +0.2, 0 and -0.5 m) were compared between locations (Goolwa Channel, Lake Alexandrina and Lake Albert) using two-factor PERMANOVA (Anderson 2001; Anderson and Ter Braak 2003) and Indicator Species Analysis (Dufrene and Legendre 1997) using the packages PCOrd version 5.12 (McCune and Mefford 2006) and PRIMER version 6.1.12 (Clarke and Gorley 2006). Bray-Curtis (1957) similarities were used to calculate the similarity matrices for all multivariate analyses and $\alpha=0.05$ for statistical analyses.

3 Results

A total of 127 taxa (including 67 exotics) were recorded for all sites between spring 2008 and autumn 2010 (Appendix 2). Functional groups, life history strategies and state and regional conservation significance for the recorded taxa are also listed in Appendix 2.

3.1 Wetlands

3.1.1 Hunter's Creek

A total of 11 taxa (including three exotics) were recorded in Hunter's Creek (Appendix 3). Despite Hunter's Creek having the lowest number of species, it also had the lowest percentage of exotics for all wetlands (Appendix 3). The middle of the creek was devoid of vegetation (except for the section between the fishway and bank, which was inundated with seawater and contained *Ruppia tuberosa*), but for the fringing plant community there were significant changes through time (Table 4). In particular there were significantly higher abundances of taxa from the terrestrial damp functional group in spring 2008, autumn 2009 and spring 2009 (Appendix 2; Appendix 4; Table 5).

3.1.2 Angas and Bremer River Mouths

A total of 28 taxa (including 15 exotics) were recorded at the mouths of the Angas and Bremer Rivers (Appendix 3). PERMANOVA (comparing the change in plant community through time and between elevations) detected significant differences in species assemblages were detected through time and between elevations and there was a significant interaction, which indicates that the changes through time were not consistent between elevations (Table 4). In the middle of the channel, the differences were predominantly due to significantly higher abundances of taxa from terrestrial damp and amphibious fluctuation responder-plastic functional groups in autumn (2009 and 2010). At the edges of the channels, differences were largely driven by significantly higher abundances of emergent and terrestrial damp taxa in spring 2008 (Appendix 2; Appendix 4; Table 6).

3.1.3 Dunn's Lagoon

A total of 57 taxa (including 30 exotics) were recorded in Dunn's Lagoon, which made this site the most species rich (total number and number of native species) wetland surveyed (Appendix 3). PERMANOVA (comparing the change in plant community through time and between elevations) detected significant differences in the floristic composition were between elevations and through time and there was a significant interaction (Table 4). General trends were taxa

from the terrestrial damp functional group tended to be in significantly higher abundances during autumn surveys, especially at the lower elevations (Appendix 2; Appendix 4; Table 7). In spring 2009, the significant indicator taxa were found from a wide range of functional groups (terrestrial dry, terrestrial damp, floodplain and amphibious fluctuation responder-plastic group) (Appendix 2; Appendix 4; Table 7). At the lowest elevation (1), differences are reflected in a greater abundance of *Chenopodium glaucum* in autumn 2009, and higher abundances of *Cotula coronopifolia* and *Senecio pterophorus* in spring 2009. At elevation 2, there were greater relative abundances of taxa from terrestrial damp functional groups in autumn 2009, and in spring 2009 grasses: *Lachnagrostis filiformis* and *Eragrostis* sp. and winter annuals: *Senecio pterophorus* and *Sonchus oleraceus* and the amphibious *Cotula coronopifolia* were present in spring 2009. In autumn 2010 taxa from the terrestrial damp functional group were in significantly higher abundances (Appendix 4; Table 7). Differences at elevation 3 reflected greater abundances of *Persicaria lapathifolia* in spring 2008 compared to greater abundances of winter annuals from terrestrial dry, terrestrial damp and amphibious functional groups in spring 2009. The plant community at elevation 4 remained relatively unchanged across time, although there were significantly higher abundances of taxa from the terrestrial damp functional group in autumn 2009. At the highest elevation (5), there were greater abundances of *Distichlis distichophylla* and *Samolus repens* in spring 2008, and winter annuals such as, *Eragrostis curvula*, *Eragrostis* sp. *Plantago coronopus*, *Reichardia tingitana* and *Sonchus oleraceus* in spring 2009.

3.1.4 Waltowa

A total of 23 taxa (including 14 exotics) were recorded in Waltowa (Appendix 3). Significant differences in floristic composition through time and between elevations were detected, and there was a significant interaction (Table 4). General trends across all elevations were that in spring 2008 significant indicator taxa were from amphibious fluctuation tolerator-emergent or floodplain functional groups, whereas in spring 2009 the significant indicator taxa were from emergent, terrestrial dry and terrestrial damp functional groups (Appendix 2; Appendix 4; Table 8). At the lowest elevation (1), there were relatively few changes through time; although there was a significantly high abundance of *Halosarcia pergranulata* in spring 2008. Similarly, for elevation 2, there were high abundances *Halosarcia pergranulata* in spring 2008; however by spring 2009 *Glyceria australis*, *Poa annua* and *Sonchus oleraceus* were significant indicators. At elevation 3, *Halosarcia pergranulata* and *Senecio runcinifolius* had higher abundances in spring 2008, but in spring 2009 *Glyceria australis* and *Sonchus oleraceus* significantly more abundant, and by autumn 2010, *Sarcocornia quinqueflora* was the most abundant species. At the highest elevation (4), *Halosarcia pergranulata* and *Senecio runcinifolius* were in high abundance in spring 2008, whereas *Lolium* spp. had a significantly greater abundance by spring 2009 (Table 8).

3.1.5 Narrung

A total of 18 taxa (including nine exotics) were recorded in Narrung Wetland (Appendix 3). There were significant differences in the plant community between elevations; however, there was no significant change through time or a significant interaction, despite the wetland being watered in spring 2009 (Table 4). At the lowest elevation (1) there were no significant indicators because this particular elevation was devoid of plants; however, there was evidence of *Ruppia* spp. (or *Lepilaena* spp.) and charophytes in March 2010 that had recruited whilst the wetland was inundated but no living plants present at the time of the survey (Appendix 4). For the remaining elevations, *Paspalum distichum* was in greater abundance at elevation 2, the salt tolerant taxa *Distichlis distichophylla*, *Frankenia pauciflora* and *Sarcocornia quinqueflora* had significantly higher abundances at elevation 3, and at the highest elevation (4), *Spergularia marina* was significantly more abundant (Table 9).

3.1.6 Goolwa Channel Drive

A total of 22 taxa (including 10 exotics) were recorded at Goolwa Channel Drive Wetland (Appendix 3). Significant changes in floristic composition were detected through time and between elevations, and there was a significant interaction (Table 4). General trends showed that in spring 2008 most of the taxa present were from the floodplain, terrestrial dry and terrestrial damp functional groups, but by autumn 2010 most taxa were present from terrestrial damp, amphibious fluctuation tolerator-emergent and emergent functional groups (Appendix 2; Appendix 4; Table 10). At the highest elevation (3) the differences were due to significantly higher abundances of *Suaeda australis* in autumn 2009 whereas, *Phragmites australis*, *Samolus repens* and *Sarcocornia quinqueflora* had greater abundances in autumn 2010 (Table 10). At elevation 2, *Lachnagrostis filiformis*, *Distichlis distichophylla*, *Plantago coronopus* and *Sonchus oleraceus* were abundant in spring 2008 and *Sarcocornia quinqueflora* and *Schoenoplectus pungens* had significantly higher abundances in autumn 2009. There were no significant indicators for spring 2009 at elevation 2, but by autumn 2010 *Paspalum distichum* and *Plantago coronopus* dominated. At the lowest elevation (1) there were significantly higher abundances of *Sonchus oleraceus* in spring 2008. In autumn 2009 there were greater abundances of the grass *Paspalum distichum*, the emergent *Bolboschoenus caldwellii* and amphibious *Schoenoplectus pungens*. In spring 2009 *Suaeda australis*, *Typha domingensis* were the most abundant species, and by autumn 2010 the most dominant species was *Atriplex prostrata* (Table 10).

3.1.7 Loveday Bay

A total of 30 taxa (including 16 exotics) were recorded in Loveday Bay (Appendix 3). The floristic composition changed significantly through time and between elevations and there was a

significant interaction (Table 4). At the lowest elevation, submergent r-selected taxa such as *Lamprothamnium macropogon* and *Ruppia tuberosa* had significantly higher abundances in spring 2008. There were no significant indicators for the intervening surveys, but by autumn 2010 taxa such as *Aster subulatus*, *Atriplex prostrata* and *Bolboschoenus caldwellii* from terrestrial damp, emergent and amphibious fluctuation tolerator-emergent functional groups, were significantly more abundant (Appendix 2; Appendix 4; Table 11). At elevation 2, *Distichlis distichophylla* had a significantly higher abundance in spring 2008, but from then on the plant community did not change significantly. At elevation (3), *Lachnagrostis filiformis*, *Bolboschoenus caldwellii* and *Distichlis distichophylla* were significantly more abundant in spring 2008, but *Schoenoplectus pungens* and *Sonchus oleraceus* were more abundant in spring 2009. At elevation 4, *Paspalum distichum* and *Sarcocornia quinqueflora* were significantly more abundant in autumn 2009, but in spring 2009, *Sonchus oleraceus* dominated, and by autumn 2010 *Suaeda australis* was more abundant. At the highest elevation (elevation 5), *Lachnagrostis filiformis* was significantly more abundant in spring 2008, but in autumn 2009, *Sarcocornia quinqueflora* dominated. In spring 2009, *Eragrostis curvula*, *Hordeum vulgare* and *Lolium* spp. were significantly more abundant and in autumn 2010 *Sonchus oleraceus* and *Paspalum distichum* had significantly higher abundances (Table 11).

3.1.8 Milang

A total of 53 taxa (including 33 exotics, the largest number in any of the wetlands surveyed) were recorded in Milang Wetland (Appendix 3). The floristic composition changed through time and between elevations, and there was a significant interaction (Table 4). Taxa within the Milang wetland were generally from terrestrial dry, terrestrial damp, amphibious fluctuation responder-plastic and amphibious fluctuation tolerator emergent functional groups (Appendix 2; Appendix 4; Table 12). At the highest elevation (1), exotic terrestrial taxa such as *Avena* spp., *Bromus diandrus* and *Medicago* spp. were dominant in spring 2008, and *Asparagus asparagoides* in spring 2009. At elevation (2), *Bromus diandrus* and *Cotula coronopifolia* were significantly more abundant in spring 2008, and in autumn 2009 the emergent species *Bolboschoenus caldwellii*, had a significantly higher abundance. By spring 2009, exotic terrestrials such as *Hordeum vulgare*, *Lolium* sp. and *Sonchus oleraceus* had a higher abundance and by autumn 2010 *Distichlis distichophylla*, *Plantago coronopus* and *Samolus repens* dominated. At elevation (3), *Lactuca saligna* and *Polypogon monspeliensis* were abundant in spring 2008, whereas *Bromus diandrus*, *Cotula coronopifolia*, *Lolium* sp. and *Sonchus oleraceus* were more abundant by spring 2009. In autumn 2010, *Cyperus gymnocaulos*, *Distichlis distichophylla*, *Enchylaena tomentosa*, *Lactuca serriola* and *Plantago coronopus* were present in greater abundances. At elevation (4), *Eragrostis* sp., *Lolium* sp. and *Polypogon monspeliensis* were more abundant in spring 2008, in autumn 2009 *Wilsonia rotundifolia* was the most abundant species and there were no significant indicators for spring 2009 and autumn 2010 (Appendix 4).

3.1.9 Poltalloch

A total of 18 taxa (including 11 exotics) were recorded in Poltalloch (Appendix 3). The floristic composition changed through time and between elevations, and there was a significant interaction (Table 4). Taxa were generally from terrestrial dry, terrestrial damp and amphibious fluctuation responder-plastic and amphibious fluctuation tolerator-emergent functional groups (Appendix 2; Appendix 4; Table 13). At elevation 1, *Chenopodium glaucum* was more abundant in spring 2008, in spring 2009 *Polygonum monspeliensis* and *Spergularia marina* were dominant and in autumn 2010, *Paspalum distichum* and *Suaeda australis* were significantly more abundant. At elevation 2, *Suaeda australis* had a significantly higher abundance in spring 2009 and *Paspalum distichum* in autumn 2010. At elevation 3, *Sarcocornia quinqueflora* were present in higher abundance in 2008, but *Distichlis distichophylla* was dominant in autumn 2009 and *Hordeum vulgare* and *Sonchus oleraceus* in spring 2009.

3.1.10 Point Sturt

A total of 34 taxa (including 23 exotics, the largest proportion of exotic taxa from the wetlands surveyed) were recorded in Point Sturt Wetland (Appendix 3). The floristic composition changed through time and between elevations and there was a significant interaction (Table 4). Taxa from this wetland were from terrestrial dry, terrestrial damp, amphibious fluctuation responder-plastic and amphibious fluctuation tolerator-emergent functional groups (Appendix 2; Appendix 4; Table 14). At elevation (1) *Chenopodium glaucum* and *Distichlis distichophylla* were present in significantly greater abundances during spring 2008, but in spring 2009 *Cotula coronopifolia*, *Eragrostis curvula*, *Reichardia tingitana*, *Sonchus oleraceus* and *Spergularia marina* were more abundant. At elevation (2), *Chenopodium album* and *Cotula coronopifolia* were more abundant in spring 2008, but by autumn 2009, *Paspalum distichum* was more abundant and in spring 2009, *Cotula coronopifolia* and *Reichardia tingitana* were in significantly higher abundances. At elevation 3, *Distichlis distichophylla* had a significantly higher abundance in spring 2008, and in autumn 2009 *Paspalum distichum* was significantly more abundant. By spring 2009 *Reichardia tingitana* was more abundant and in autumn 2010, *Aster subulatus*, *Enchylaena tomentosa* and *Sarcocornia quinqueflora* were more abundant (Table 14).

3.1.11 Teringie

A total of 30 taxa (including 18 exotics) were recorded in Teringie (Appendix 3). The floristic composition changed through time and between elevations, and there was a significant interaction (Table 4). Taxa from the terrestrial dry functional group were in significantly higher abundances in spring 2009, but in the following surveys taxa in high abundances were from a wider range of functional groups (terrestrial dry, terrestrial damp, amphibious fluctuation

tolerator-emergent) (Appendix 2; Appendix 4; Table 15). At the highest elevation (elevation 1), *Distichlis distichophylla* and *Sarcocornia quinqueflora* had significantly higher abundances in autumn 2009; otherwise there were minimal changes in floristic composition through time. For elevation 2, *Senecio* sp. was more abundant in spring 2008; otherwise there were no significant changes in floristic composition through time. For the lowest elevation (3), *Frankenia pauciflora* and *Senecio* sp. had significantly higher abundances in spring 2008, but by the following spring (2009), *Eragrostis curvula* and *Polypogon monspeliensis* were the most abundant. For elevation 4, *Hordeum vulgare* was the most abundant species in spring 2008, but by *Avena* spp., *Lolium* spp. and *Medicago* spp. were most abundant in spring 2009 and in autumn 2010, *Enchylaena tomentosa* was the most abundant species (Table 15).

Table 4: PERMANOVA *Pseudo F*-statistic results comparing time and elevation for each wetland.

| Wetland | Factor | df | F | P |
|-------------------------------|------------------|---------|-------|--------|
| Hunter's Creek | Time | 3, 48 | 8.78 | <0.001 |
| Angas and Bremer River Mouths | Elevation | 1,48 | 9.88 | <0.001 |
| | Time | 3, 48 | 1.92 | 0.018 |
| | Elevation × Time | 3, 48 | 2.24 | 0.003 |
| Dunn's Lagoon | Elevation | 4, 239 | 40.54 | <0.001 |
| | Time | 3, 239 | 7.01 | 0.11 |
| | Elevation × Time | 12, 239 | 3.82 | 0.12 |
| Waltowa | Elevation | 3, 96 | 10.83 | <0.001 |
| | Time | 3, 96 | 22.59 | <0.001 |
| | Elevation × Time | 9, 96 | 2.89 | <0.001 |
| Narrung | Elevation | 3, 144 | 30.57 | <0.001 |
| | Time | 3, 144 | 1.1 | 0.34 |
| | Elevation × Time | 9, 144 | 0.89 | 0.591 |
| Goolwa Channel Drive | Elevation | 2, 108 | 37.71 | <0.001 |
| | Time | 3, 108 | 5.85 | <0.001 |
| | Elevation × Time | 6, 108 | 4.19 | <0.001 |
| Loveday Bay | Elevation | 4, 180 | 5.68 | <0.001 |
| | Time | 3, 180 | 8.09 | <0.001 |
| | Elevation × Time | 12, 180 | 2.43 | <0.001 |
| Milang | Elevation | 3, 144 | 9.14 | <0.001 |
| | Time | 3, 144 | 4.09 | <0.001 |
| | Elevation × Time | 9, 144 | 2.20 | <0.001 |
| Poltalloch | Elevation | 2, 72 | 9.30 | <0.001 |
| | Time | 3, 72 | 9.51 | <0.001 |
| | Elevation × Time | 6, 72 | 2.83 | <0.001 |
| Point Sturt | Elevation | 2, 72 | 7.89 | <0.001 |
| | Time | 3, 72 | 15.98 | <0.001 |
| | Elevation × Time | 6, 72 | 3.62 | <0.001 |
| Teringie | Elevation | 3, 192 | 21.58 | <0.001 |
| | Time | 3, 192 | 7.36 | <0.001 |
| | Elevation × Time | 9, 192 | 7.40 | <0.001 |

Table 5: List of significant indicator species (Dufrene and Legendre 1997) for Hunter's Creek between elevations from spring 2008 to autumn 2010.

| Elevation | Time | | | |
|-----------|---------------------------|----------------------------|----------------------------------|-------------|
| | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Edge | <i>Sonchus oleraceus*</i> | <i>Paspalum distichum*</i> | <i>Distichlis distichophylla</i> | |
| Middle | bare | bare | bare | bare |

Table 6: List of significant indicator species (Dufrene and Legendre 1997) for Angus and Bremer River Mouths between elevations from spring 2008 to autumn 2010.

| Elevation | Time | | | |
|-----------|---|---|-------------|---|
| | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Edge | <i>Berula erecta</i> , <i>Sonchus oleraceus*</i> , <i>Typha domingensis</i> | | | |
| Middle | | <i>Aster subulatus*</i> , <i>Atriplex prostrata*</i> | | <i>Chenopodium album*</i> , <i>Paspalum distichum*</i> , <i>Persicaria lapathifolia</i> |

Table 7: List of significant indicator species (Dufrene and Legendre 1997) for Dunn's Lagoon between elevations from spring 2008 to autumn 2010.

| Elevation | Time | | | |
|------------------------|-------------|---|--|--|
| | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Lowest ↑ Highest | 1 | | <i>Chenopodium glaucum*</i> | <i>Cotula coronopifolia*</i> , <i>Senecio pterophorus*</i> |
| | 2 | | <i>Aster subulatus*</i> , <i>Chenopodium glaucum*</i> | <i>Lachnagrostis filiformis</i> , <i>Eragrostis sp.</i> , <i>Senecio pterophorus*</i> , <i>Sonchus oleraceus*</i> , <i>Cotula coronopifolia*</i> |
| | 3 | <i>Persicaria lapathifolia</i> | | <i>Reichardia tingitana*</i> , <i>Sonchus oleraceus*</i> , <i>Cotula coronopifolia*</i> |
| | 4 | | <i>Paspalum distichum*</i> | |
| | 5 | <i>Distichlis distichophylla</i> , <i>Samolus repens</i> | | <i>Eragrostis curvula*</i> , <i>Eragrostis sp.</i> , <i>Plantago coronopus*</i> , <i>Reichardia tingitana*</i> , <i>Sonchus oleraceus*</i> |

Table 8: List of significant indicator species (Dufrene and Legendre 1997) for Angus and Bremer River Mouths Wetlands between elevations from spring 2008 to autumn 2010.

| Elevation | | Time | | | |
|------------------------|---|--|-------------|--|-----------------------------------|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Highest ↑ Lowest | 1 | <i>Halosarcia pergranulata</i> | | | |
| | 2 | <i>Halosarcia pergranulata</i> | | <i>Glyceria australis</i> , <i>Poa annua</i> *, <i>Sonchus oleraceus</i> * | |
| | 3 | <i>Halosarcia pergranulata</i> , <i>Senecio runcinifolius</i> * | | <i>Glyceria australis</i> , <i>Sonchus oleraceus</i> * | <i>Sarcocornia quinqueloflora</i> |
| | 4 | <i>Halosarcia pergranulata</i> , <i>Senecio runcinifolius</i> * | | <i>Lolium</i> spp.* | |

Table 9: List of significant indicator species (Dufrene and Legendre 1997) for Narrung Wetland between elevations.

| Elevation | | |
|------------------------|---|--|
| Highest ↑ Lowest | 1 | |
| | 2 | <i>Paspalum distichum</i> * |
| | 3 | <i>Distichlis distichophylla</i> , <i>Frankenia pauciflora</i> , <i>Sarcocornia quinqueloflora</i> |
| | 4 | <i>Spergularia marina</i> |
| | 5 | bare |

Table 10: List of significant indicator species (Dufrene and Legendre 1997) for Goolwa Channel Drive Wetland between elevations from spring 2008 to autumn 2010.

| Elevation | | Time | | | |
|------------------------|---|---|--|---|---|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Highest ↑ Lowest | 1 | <i>Sonchus oleraceus</i> * | <i>Paspalum distichum</i> *, <i>Bolboschoenus caldwellii</i> , <i>Schoenoplectus pungens</i> | <i>Suaeda australis</i> , <i>Typha domingensis</i> | <i>Atriplex prostrata</i> * |
| | 2 | <i>Lachnagrostis filiformis</i> , <i>Distichlis distichophylla</i> , <i>Plantago coronopus</i> *, <i>Sonchus oleraceus</i> * | <i>Sarcocornia quinqueloflora</i> , <i>Schoenoplectus pungens</i> | | <i>Paspalum distichum</i> *, <i>Plantago coronopus</i> * |
| | 3 | | <i>Suaeda australis</i> | | <i>Phragmites australis</i> , <i>Samolus repens</i> , <i>Sarcocornia quinqueloflora</i> |

Table 11: List of significant indicator species (Dufrene and Legendre 1997) for Loveday Bay Wetland between elevations from spring 2008 to autumn 2010.

| Elevation | | Time | | | |
|-------------|---|--|---|---|--|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Lowest ↑ | 1 | <i>Lamprothamnium macropogon</i> , <i>Ruppia tuberosa</i> | | | <i>Aster subulatus</i> *, <i>Atriplex prostrata</i> *, <i>Bolboschoenus caldwellii</i> |
| | 2 | <i>Distichlis distichophylla</i> | | | |
| | 3 | <i>Lachnagrostis filiformis</i> , <i>Bolboschoenus caldwellii</i> , <i>Distichlis distichophylla</i> | | <i>Schoenoplectus pungens</i> , <i>Sonchus oleraceus</i> * | |
| | 4 | | <i>Paspalum distichum</i> *, <i>Sarcocornia quinqueflora</i> | <i>Sonchus oleraceus</i> * | <i>Suaeda australis</i> |
| Highest | 5 | <i>Lachnagrostis filiformis</i> | <i>Sarcocornia quinqueflora</i> | <i>Eragrostis curvula</i> *, <i>Hordeum vulgare</i> *, <i>Lolium spp.</i> * | <i>Sonchus oleraceus</i> *, <i>Paspalum distichum</i> * |

Table 12: List of significant indicator species (Dufrene and Legendre 1997) for Milang Wetland between elevations from spring 2008 to autumn 2010.

| Elevation | | Time | | | |
|--------------|---|--|---------------------------------|--|--|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Highest ↓ | 1 | <i>Avena spp.</i> *, <i>Bromus diandrus</i> *, <i>Medicago spp.</i> * | | <i>Asparagus asparagoides</i> * | |
| | 2 | <i>Bromus diandrus</i> *, <i>Cotula coronopifolia</i> * | <i>Bolboschoenus caldwellii</i> | <i>Hordeum vulgare</i> *, <i>Lolium spp.</i> *, <i>Sonchus oleraceus</i> * | <i>Distichlis distichophylla</i> , <i>Plantago coronopus</i> *, <i>Samolus repens</i> |
| | 3 | <i>Lactuca saligna</i> *, <i>Polypogon monspeliensis</i> * | | <i>Bromus diandrus</i> *, <i>Cotula coronopifolia</i> *, <i>Lolium spp.</i> *, <i>Sonchus oleraceus</i> * | <i>Cyperus gymnocaulos</i> , <i>Distichlis distichophylla</i> , <i>Enchylaena tomentosa</i> , <i>Lactuca serriola</i> *, <i>Plantago coronopus</i> * |
| | 4 | <i>Eragrostis sp.</i> , <i>Lolium spp.</i> *, <i>Polypogon monspeliensis</i> * | <i>Wilsonia rotundifolia</i> | | |

Table 13: List of significant indicator species (Dufrene and Legendre 1997) for Pottaloch Wetland between elevations from spring 2008 to autumn 2010.

| Elevation | | Time | | | |
|------------------------|---|---------------------------------|----------------------------------|--|---|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Highest ↑ Lowest | 1 | <i>Chenopodium glaucum</i> * | | <i>Polypogon monspeliensis</i> *, <i>Spergularia marina</i> * | <i>Paspalum distichum</i> *, <i>Suaeda australis</i> |
| | 2 | | | <i>Suaeda australis</i> | <i>Paspalum distichum</i> * |
| | 3 | <i>Sarcocornia quinqueflora</i> | <i>Distichlis distichophylla</i> | <i>Hordeum vulgare</i> *, <i>Sonchus oleraceus</i> * | |

Table 14: List of significant indicator species (Dufrene and Legendre 1997) for Point Sturt Wetland between elevations from spring 2008 to autumn 2010 (*denotes exotic taxon)

| Elevation | | Time | | | |
|-------------|---|---|----------------------------|---|--|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Lowest ↑ | 1 | <i>Chenopodium glaucum*</i> <i>Distichlis distichophylla</i> | | <i>Cotula coronopifolia*</i> , <i>Eragrostis curvula*</i> , <i>Reichardia tingitana*</i> , <i>Sonchus oleraceus*</i> , <i>Spergularia marina*</i> | |
| | 2 | <i>Chenopodium album*</i> <i>Cotula coronopifolia*</i> | <i>Paspalum distichum*</i> | <i>Cotula coronopifolia*</i> , <i>Reichardia tingitana*</i> | |
| Highest | 3 | <i>Distichlis distichophylla</i> | <i>Paspalum distichum*</i> | <i>Reichardia tingitana*</i> | <i>Aster subulatus</i> , <i>Enchylaena tomentosa</i> , <i>Sarcocornia quinqueloflora</i> |

Table 15: List of significant indicator species (Dufrene and Legendre 1997) for Teringie Wetland between elevations from spring 2008 to autumn 2010 (*denotes exotic taxon)

| Elevation | | Time | | | |
|-------------|---|---|---|---|-----------------------------|
| | | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 |
| Lowest ↑ | 1 | | <i>Distichlis distichophylla</i> <i>Sarcocornia quinqueloflora</i> | | |
| | 2 | <i>Senecio</i> sp. | | | |
| Highest | 3 | <i>Frankenia pauciflora</i> , <i>Senecio</i> sp. | | <i>Eragrostis curvula*</i> <i>Polypogon monspeliensis*</i> | |
| | 4 | <i>Hordeum vulgare*</i> | <i>Avena</i> spp* <i>Lolium</i> spp.* <i>Medicago</i> spp.* | | <i>Enchylaena tomentosa</i> |

3.2 Lakeshores

3.2.1 Changes between locations (Lake Alexandrina, Lake Albert and Goolwa Channel) from spring 2009 to autumn 2010 at each elevation

3.2.1.1 +0.8 m AHD

The floristic composition changed through time and between locations, and there was a significant interaction (Table 16). At the highest elevation (+0.8 m AHD) there were significantly higher abundances of species from the terrestrial dry and terrestrial damp functional groups in Lake Alexandrina in spring 2009, but no significant indicator species in autumn 2010 (Appendix 2; Appendix 4; Table 17). Similarly for Lake Albert, taxa from the terrestrial dry functional group were significantly abundant during spring 2009, but no significant indicator species for autumn 2010 (Appendix 2; Appendix 4; Table 17). In Goolwa Channel, taxa from a wider range of functional groups (amphibious fluctuation tolerant emergent, amphibious fluctuation tolerator woody and emergent) were significantly more abundant in spring 2009, whereas the emergent *Phragmites australis* dominated in autumn 2010 (Appendix 2; Appendix 4; Table 17).

3.2.1.2 +0.6 m AHD

There were significant differences in floristic composition between location and time, but there was no significant interaction (Table 16). In spring 2009, *Hypochoeris radicata* (terrestrial dry functional group) was significantly more abundant in Lake Alexandrina, but there were no significant indicators in autumn 2010 (Appendix 2; Appendix 4; Table 17). Similarly in Lake Albert, taxa from the terrestrial dry and terrestrial damp functional groups were significantly more abundant in spring 2009. In autumn 2010, taxa from the terrestrial dry, terrestrial damp, amphibious fluctuation tolerator woody functional groups were significantly more abundant (Appendix 2; Appendix 4; Table 17). In the Goolwa Channel, *Muehlenbeckia florulenta* (amphibious fluctuation tolerator woody) and *Phragmites australis* (emergent) were significant indicators in spring 2009, whereas *Bolboschoenus caldwellii* (emergent) were significantly abundant during autumn 2010 (Appendix 2; Appendix 4; Table 17).

3.2.1.3 +0.4 m AHD

There were significant differences in floristic composition between location and time and a significant interaction (Table 16). In Lake Alexandrina taxa from terrestrial dry, terrestrial damp and amphibious fluctuation responder-plastic functional groups were significantly more abundant in Spring 2009 while in autumn 2010 taxa from the same terrestrial dry, terrestrial damp and amphibious fluctuation tolerant emergent functional groups were significant indicators (Appendix 2; Appendix 4; Table 17). For Lake Albert, similar trends were observed with the significant indicators present from the terrestrial dry, terrestrial damp and amphibious fluctuation tolerator emergent functional groups in spring 2009 and autumn 2010. In Goolwa Channel, *Muehlenbeckia florulenta* was significantly more abundant in spring 2009 and in autumn 2010 *Juncus usitatus* (amphibious fluctuation tolerator-emergent) and *Typha domingensis* (emergent) were significant indicators (Appendix 2; Appendix 4; Table 17).

3.2.1.4 +0.2 m AHD

There were significant differences in floristic composition between location and time and there was also a significant interaction (Table 16). In Lake Alexandrina taxa from terrestrial dry, terrestrial damp and amphibious fluctuation tolerator emergent functional groups were significantly more abundant in spring 2009 and autumn 2010 (Appendix 2; Appendix 4; Table 17). The same trends were evident in Lake Albert (Appendix 2; Appendix 4; Table 17). In Goolwa Channel there were no significant indicators at this elevation in spring 2009, but in autumn 2010 *Atriplex prostrata* (terrestrial damp) and the emergent species *Schoenoplectus validus* and *Typha domingensis* were significantly more abundant (Table 17).

3.2.1.5 0 m AHD

There were significant differences in floristic composition between location and time and there was also a significant interaction (Table 16). In Lake Alexandrina taxa from a wide range of functional groups (terrestrial dry, terrestrial damp, floodplain and amphibious fluctuation tolerator emergent) were significantly more abundant in spring 2009, while in autumn 2010 the sedges *Juncus kraussii* and *Juncus usitatus* (amphibious fluctuation tolerator emergent) were significantly indicators (Appendix 2; Appendix 4; Table 17). Similarly there were significant indicators in Lake Albert, from a wide range of functional groups (terrestrial dry, terrestrial damp, amphibious fluctuation responder-plastic and amphibious fluctuation tolerator emergent) in spring 2009, only *Paspalum distichum* was significantly more abundant in autumn 2010 (Appendix 2; Appendix 4; Table 17). There were no significant indicator species for Goolwa Channel in spring 2009, although the emergent species *Schoenoplectus validus* was significantly more abundant in autumn 2010 (Table 17).

3.2.1.6 -0.5 m AHD

Significant differences in floristic composition were detected between location and time and there was also a significant interaction (Table 16). In Lake Alexandrina a large number of taxa from a wide range of functional groups (floodplain, terrestrial dry, terrestrial damp, amphibious floodplain responder plastic, amphibious fluctuation tolerator low-growing and amphibious fluctuation tolerator emergent) were significantly more abundant in spring 2009. In autumn 2010 there were fewer significant indicators, from fewer functional groups (terrestrial dry, terrestrial damp, amphibious fluctuation tolerator emergent) (Appendix 2; Appendix 4; Table 17). In Lake Albert the significant indicators in spring 2009 were from the terrestrial damp, amphibious fluctuation tolerator emergent and amphibious fluctuation tolerator woody (*Melaleuca balmaturorum*) groups. In autumn the terrestrial damp species, *Chenopodium album* was significantly more abundant (Appendix 2; Appendix 4; Table 17). In Goolwa Channel there were no significant indicators in spring 2009, but *Myriophyllum salsugineum* and *Potamogeton pectinatus* (submergent k-selected) were significantly more abundant in autumn 2010 (Table 17).

Table 16: PERMANOVA Pseudo-F-statistic results through time and location (Lake Alexandrina, Lake Albert and Goolwa channel region) per elevation.

| Elevation | Factor | df | F | P |
|------------------|-----------------|-----------|----------|----------|
| +0.8 m AHD | Location | 2, 143 | 20.56 | <0.001 |
| | Time | 1, 143 | 5.06 | <0.001 |
| | Location × Time | 2, 143 | 1.92 | 0.041 |
| +0.6 m AHD | Location | 2, 143 | 14.77 | <0.001 |
| | Time | 1, 143 | 2.72 | 0.022 |
| | Location × Time | 2, 143 | 1.38 | 0.178 |
| +0.4 m AHD | Location | 2, 143 | 9.48 | <0.001 |
| | Time | 1, 143 | 3.62 | 0.005 |
| | Location × Time | 2, 143 | 2.65 | 0.004 |
| +0.2 m AHD | Location | 2, 143 | 10.45 | <0.001 |
| | Time | 1, 143 | 3.71 | 0.003 |
| | Location × Time | 2, 143 | 2.32 | 0.009 |
| 0 m AHD | Location | 2, 143 | 12.54 | <0.001 |
| | Time | 1, 143 | 9.23 | <0.001 |
| | Location × Time | 2, 143 | 3.89 | <0.001 |
| -0.5 m AHD | Location | 2, 143 | 6.11 | <0.001 |
| | Time | 1, 143 | 6.26 | <0.001 |
| | Location × Time | 2, 143 | 4.77 | <0.001 |

Table 17: Significant indicator species (Dufrene and Legendre 1997) for locations (Lake Alexandrina, Lake Albert and Goolwa Channel) from spring 2009 to autumn 2010, at each elevation (*denotes exotic taxon).

| Location | Elevation | | | | | | | | | | | |
|-------------------------|---|-----------------------------|---|---|---|---|--|---|--|---|--|---|
| | +0.8 m AHD | | +0.6 m AHD | | +0.4 m AHD | | +0.2 m AHD | | 0 m AHD | | -0.5 m AHD | |
| | Spring 2009 | Autumn 2010 | Spring 2009 | Autumn 2010 | Spring 2009 | Autumn 2010 | Spring 2009 | Autumn 2010 | Spring 2009 | Autumn 2010 | Spring 2009 | Autumn 2010 |
| Lake Alexandrina | <i>Hordeum vulgare</i> *, <i>Lactuca serriola</i> *, <i>Sonchus oleraceus</i> * | | <i>Hypochoeris radicata</i> * | | <i>Hordeum vulgare</i> *, <i>Persicaria lapathifolia</i> , <i>Senecio pterophorus</i> *, <i>Sonchus asper</i> * | | <i>Isolepis nodosa</i> , <i>Paspalum distichum</i> *, <i>Plantago coronopus</i> * | <i>Conyza bonariensis</i> *, <i>Cyperus gymnocaulos</i> , <i>Hordeum vulgare</i> * | <i>Aster subulatus</i> *, <i>Juncus kraussii</i> , <i>Plantago coronopus</i> * | <i>Centaurea calcitrapa</i> *, <i>Eragrostis curvula</i> *, <i>Hordeum vulgare</i> *, <i>Hypochoeris radicata</i> *, <i>Lachnagrostis filiformis</i> , <i>Lactuca serriola</i> *, <i>Polypogon monspeliensis</i> *, <i>Senecio pterophorus</i> *, <i>Trifolium spp.</i> * | <i>Lachnagrostis filiformis</i> , <i>Aster subulatus</i> *, <i>Centaureum tenuiflorum</i> *, <i>Cotula coronopifolia</i> *, <i>Eragrostis curvula</i> *, <i>Hordeum vulgare</i> *, <i>Isolepis platycarpa</i> , <i>Lobelia alata</i> , <i>Lolium spp.</i> , <i>Plantago coronopus</i> *, <i>Polypogon monspeliensis</i> *, <i>Sonchus oleraceus</i> * | <i>Centaurea calcitrapa</i> *, <i>Chenopodium glaucum</i> *, <i>Paspalum distichum</i> *, <i>Senecio pterophorus</i> *, <i>Suaeda australis</i> |
| Lake Albert | <i>Avena spp.</i> *, <i>Bromus diandrus</i> *, <i>Pennisetum clandestinum</i> *, <i>Trifolium spp.</i> *, <i>Vicia sativa</i> * | | <i>Avena spp.</i> *, <i>Bromus diandrus</i> *, <i>Lolium spp.</i> *, <i>Pennisetum clandestinum</i> *, <i>Sonchus oleraceus</i> * | <i>Paspalum distichum</i> *, <i>Reichardia tingitana</i> *, <i>Suaeda australis</i> | <i>Bromus diandrus</i> *, <i>Lolium spp.</i> *, <i>Medicago spp.</i> *, <i>Pennisetum clandestinum</i> *, <i>Polypogon monspeliensis</i> *, <i>Sonchus oleraceus</i> *, <i>Suaeda australis</i> , | <i>Reichardia tingitana</i> *, <i>Sarcocornia quinqueflora</i> , <i>Scabiosa atropurpurea</i> *, <i>Spergularia marina</i> , | <i>Avena spp.</i> *, <i>Bromus diandrus</i> *, <i>Lolium spp.</i> *, <i>Paspalum distichum</i> *, <i>Pennisetum clandestinum</i> *, <i>Polypogon monspeliensis</i> *, <i>Sonchus oleraceus</i> * | <i>Eragrostis australasica</i> , <i>Reichardia tingitana</i> *, <i>Sarcocornia quinqueflora</i> , | <i>Avena spp.</i> *, <i>Bromus diandrus</i> *, <i>Cotula coronopifolia</i> *, <i>Euphorbia terracina</i> *, <i>Lolium spp.</i> *, <i>Medicago spp.</i> *, <i>Sarcocornia quinqueflora</i> , <i>Sonchus oleraceus</i> *, | <i>Cyperus gymnocaulos</i> , <i>Isolepis nodosa</i> , <i>Melaleuca halimaturorum</i> , <i>Puccinellia spp.</i> | <i>Chenopodium album</i> * | |
| Goolwa Channel | <i>Calystegia sepium</i> , <i>Muehlenbeckia florulenta</i> , <i>Typha domingensis</i> | <i>Phragmites australis</i> | <i>Muehlenbeckia florulenta</i> , <i>Phragmites australis</i> | <i>Muehlenbeckia florulenta</i> , | <i>Juncus usitatus</i> , <i>Typha domingensis</i> | | <i>Atriplex prostrata</i> *, <i>Schoenoplectus validus</i> , <i>Typha domingensis</i> | | <i>Schoenoplectus validus</i> , | | <i>Myriophyllum salicagineum</i> , <i>Potamogeton pectinatus</i> | |

3.2.2 Changes within locations (Lake Alexandrina, Lake Albert and Goolwa Channel) between elevations from spring 2008 to autumn 2010

3.2.2.1 Lake Alexandrina

Significant differences in the plant community in Lake Alexandrina were detected through time and between elevations and there was a significant interaction (Table 18). At the highest elevation (+0.8 m AHD) terrestrial dry annuals and *Salix babylonica* (emergent) were significantly more abundant in spring 2008 but there were no significant indicators in the following autumn and spring 2009. By autumn 2010 *Pennisetum clandestinum* was significantly more abundant (Appendix 2; Appendix 4; Table 19). At +0.6 m AHD *Melilotus indica* was significantly more abundant in spring 2008 and *Bromus diandrus* in spring 2009; however, there were no significant indicators in autumn 2009 and 2010 (Table 19). At +0.4 m AHD, there were significant indicators from a range of functional groups (floodplain, terrestrial damp, amphibious fluctuation responder-plastic and amphibious fluctuation tolerator emergent) in spring 2008, but no further significant indicators until autumn 2010 when *Isolepis nodosa* was significantly abundant (Appendix 2; Appendix 4; Table 19). At 0 m AHD exotic terrestrial dry taxa, exotic terrestrial damp taxa, *Isolepis* sp. (amphibious fluctuation tolerator–low growing) and *Schoenoplectus validus* (emergent) were significantly more abundant in spring 2008 (Appendix 4). In autumn 2009 *Polygonum aviculare* was significantly more abundant, and terrestrial damp taxa were significant indicators in spring 2009. In autumn 2010 *Juncus usitatus* and *Paspalum distichum* were significantly more abundant (Appendix 2; Appendix 4; Table 19). At the lowest elevation (-0.5 m AHD) the submergent r-selected *Ruppia tuberosa* was a significant indicator in spring 2008, but by autumn 2009 *Eragrostis* sp. was significantly more abundant (Appendix 4). In the following spring (2009), exotics from a wide range of functional groups (floodplain, terrestrial damp, terrestrial dry, amphibious fluctuation tolerator-low growing) were significantly more abundant. In autumn 2010 taxa from emergent, amphibious fluctuation tolerator-low growing, terrestrial damp and amphibious fluctuation tolerator emergent groups were significant indicators (Appendix 2; Appendix 4; Table 19).

3.2.2.2 Lake Albert

Significant differences in the plant community in Lake Albert were detected through time and between elevations, and there was a significant interaction (Table 18). At the highest elevation (+0.8 m AHD) *Distichlis distichophylla* was significantly more abundant in spring 2008, and terrestrial dry taxa were significant indicators in the following spring (2009). In autumn 2010 *Pennisetum clandestinum* was significantly more abundant (Appendix 2; Appendix 4; Table 19). At +0.6 m AHD *Melilotus indica* was significantly more abundant in spring 2008, and *Bromus diandrus* in spring 2009 (Appendix 2; Appendix 4; Table 19). There were minimal changes through time at

+0.4 m AHD, although *Polygogon monspeliensis* and *Schoenoplectus validus* were significantly more abundant in spring 2008 (Table 19). At 0 m AHD there were no significant indicators in spring 2008 and autumn 2009. By spring 2009, *Sonchus oleraceus* was significantly more abundant and by autumn 2010 *Aster subulatus* and *Paspalum distichum* were significant indicators (Table 19). At the lowest elevation (-0.5 m AHD) there were no significant indicators in spring 2008 or autumn 2009 but by spring 2009 *Cotula coronopifolia*, *Isolepis nodosa*, *Puccinellia* spp. and *Melaleuca balmaturorum* seedlings were significantly more abundant. In autumn 2010, *Chenopodium album* (terrestrial damp) was a significant indicator (Appendix 2; Appendix 4; Table 19).

3.2.2.3 Goolwa Channel

Significant differences in the floristic composition of Goolwa Channel were detected through time and between elevations, and there was also a significant interaction (Table 18). At the highest elevation (+0.8 m AHD) *Calystegia sepium* (amphibious fluctuation tolerator-emergent species) was significantly more abundant in spring 2008, and there were no significant indicators in autumn 2009 and spring 2009. By autumn 2010, however, *Paspalum distichum* was significantly more abundant. At +0.6 m AHD, there were no significant indicator species in the first year, but by spring 2009 *Cyperus gymnocaulos* and *Isolepis nodosa* (amphibious fluctuation tolerator emergent species) were significantly more abundant and the emergent species *Bolboschoenus caldwellii* was significantly more abundant in autumn 2010 (Appendix 2; Appendix 4; Table 19). In spring 2008 floodplain, terrestrial damp, amphibious fluctuation tolerator emergent taxa were significantly more abundant at +0.4 m AHD. In autumn 2009, *Chenopodium glaucum* was significantly more abundant and there were no significant indicators in spring 2009. In autumn 2010 *Juncus usitatus* (amphibious fluctuation tolerator-emergent) and *Typha domingensis* (emergent) were significantly more abundant (Appendix 2; Appendix 4; Table 19). At 0 m AHD *Cotula coronopifolia* had a significantly higher abundance in spring 2008, but there were no significant indicators in the subsequent surveys. At the lowest elevation (-0.5 m AHD) there were no significant indicators until autumn 2010, when the submergent k-selected species: *Myriophyllum salsugineum* was significantly more abundant (Appendix 4; Table 19).

Table 18: PERMANOVA Pseudo-F-statistic results comparing time (spring 2008, autumn 2009, spring 2009 and autumn 2010) and location (Lake Alexandrina, Lake Albert and Goolwa channel region) for each elevation.

| Location | Factor | df | F | P |
|------------------|------------------|-----------|----------|----------|
| Lake Alexandrina | Elevation | 4, 359 | 13.22 | <0.001 |
| | Time | 3, 359 | 7.56 | <0.001 |
| | elevation × Time | 12, 359 | 2.06 | <0.001 |
| Lake Albert | Location | 4, 119 | 10.75 | <0.001 |
| | Time | 3, 119 | 22.97 | <0.001 |
| | Location × Time | 12, 119 | 4.15 | <0.001 |
| Goolwa Channel | Elevation | 4, 119 | 21.81 | <0.001 |
| | Time | 3, 119 | 4.4 | <0.001 |
| | elevation × Time | 12, 119 | 3.5 | <0.001 |

Table 19: Significant indicator species (Dufrene and Legendre 1997) within locations (Lake Alexandrina, Lake Albert and Goolwa Channel) and between elevations from spring 2008 to autumn 2010 (*denotes exotic taxon).

| Elevation | Location | | | | | | | | | | | | | | | | |
|------------|---|------------------------------|---|---|--|-------------|---|--|-------------|-------------|-------------|----------------------------|---|------------------------------|--|--|----------------------------------|
| | Lake Alexandrina | | | | | | Lake Albert | | | | | | Goolwa Channel | | | | |
| | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 | Spring 2008 | Autumn 2009 | Spring 2009 | Autumn 2010 | |
| +0.8 m AHD | <i>Brassica rapa</i> *, <i>Senecio</i> sp., <i>Vicia sativa</i> *, <i>Salix babylonica</i> * | | | <i>Pennisetum clandestinum</i> * | <i>Distichlis distichophylla</i> | | <i>Avena</i> spp.*, <i>Lolium</i> spp.*, <i>Trifolium</i> spp.*, <i>Vicia sativa</i> * | <i>Pennisetum clandestinum</i> * | | | | <i>Calyptegia sepium</i> , | | | | <i>Paspalum distichum</i> * | |
| +0.6 m AHD | <i>Melilotus indica</i> * | | <i>Bromus diandrus</i> * | | <i>Melilotus indica</i> * | | | | | | | | | | <i>Cyperus gymnocaulos</i> , <i>Isolepis nodosa</i> | <i>Bolboschoenus caldwelii</i> | |
| +0.4 m AHD | <i>Cotula coronopifolia</i> *, <i>Cyperus gymnocaulos</i> , <i>Holcus lanatus</i> *, <i>Pseudognaphalium luteo-album</i> | | | <i>Isolepis nodosa</i> | <i>Polygonum monspeliensis</i> *, <i>Schoenoplectus validus</i> | | | | | | | | <i>Lachnagrostis filiformis</i> , <i>Aster subulatus</i> *, <i>Juncus kraussii</i> *, <i>Epilobium pallidiflorum</i> , <i>Sonchus oleraceus</i> * | <i>Chenopodium glaucum</i> * | | <i>Juncus usitatus</i> , <i>Typha domingensis</i> | |
| 0 m AHD | <i>Aster subulatus</i> *, <i>Eragrostis curvula</i> *, <i>Polygonum monspeliensis</i> *, <i>Isolepis</i> sp., <i>Schoenoplectus validus</i> , | <i>Polygonum aviculare</i> * | <i>Centaura calcitrapa</i> *, <i>Sonchus oleraceus</i> * | <i>Juncus usitatus</i> , <i>Paspalum distichum</i> * | | | <i>Sonchus oleraceus</i> * | <i>Aster subulatus</i> *, <i>Paspalum distichum</i> * | | | | | | | | | |
| -0.5 m AHD | <i>Ruppia tuberosa</i> | <i>Eragrostis</i> sp. | <i>Lachnagrostis filiformis</i> , <i>Centaurium tenuiflorum</i> *, <i>Hordeum vulgare</i> *, <i>Isolepis platycarpa</i> , <i>Lobelia alata</i> , <i>Lolium</i> spp.*, <i>Plantago coronopus</i> * | <i>Bolboschoenus caldwelii</i> , <i>Triglochin striatum</i> , <i>Spergularia marina</i> *, <i>Suaeda australis</i> | | | <i>Cotula coronopifolia</i> *, <i>Isolepis nodosa</i> , <i>Puccinellia</i> spp., <i>Melaleuca halmaturorum</i> | <i>Chenopodium album</i> * | | | | | | | | | <i>Myriophyllum salisugineum</i> |

4 Discussion and management implications

4.1 Wetlands

The sustained decline in water levels downstream of Lock 1 between 2007 and 2010 has resulted in permanent wetlands that were historically connected to Lakes Alexandrina and Albert becoming disconnected and drying. Prior to 2007 the plant community in the Lower Lakes wetlands was a diverse assemblage of submergent, amphibious and emergent taxa (Holt *et al.* 2005; Nicol *et al.* 2006). By spring 2008, all submergent species except *Lamprothamnium macropogon* (Loveday Bay), *Ruppia tuberosa* (Loveday Bay and Hunters Creek) and *Ruppia polycarpa* (Milang) had been lost from the system (Marsland and Nicol 2009). In addition to submergent species, a total of 33 species of emergent, floating and amphibious taxa were lost from wetlands between 2005 and 2008 (Holt *et al.* 2005; Nicol *et al.* 2006; Marsland and Nicol 2009). Whilst there have been no further species losses in 2009-10, wetlands have remained dry and there has been a trend of terrestrial species that were historically confined to high elevations colonising lower elevations (*sensu* Brock and Casanova 1997; Nicol *et al.* 2003).

Similar to 2008-09 the changes in floristic composition through time were generally due to changes in abundance of winter annuals; however, terrestrial taxa have colonised lower elevations than in 2008-09 (Marsland and Nicol 2009). Wetlands subjected to prolonged drawdown are generally invaded by terrestrial taxa because conditions for the growth and survival of wetland taxa are unsuitable (Nicol *et al.* 2003, Nicol 2010). The majority of terrestrial taxa that have colonised the Lower Lakes wetlands are agricultural weeds, typically pasture grasses and legumes (Cunningham *et al.* 1981; Jessop and Tolken 1986; Dashorst and Jessop 1998; Jessop *et al.* 2006). Invasion of agricultural weeds is probably due to the surrounding land use (predominantly agricultural and urban); furthermore, Nicol and Ward (2010 b) reported that the sediment seed bank of Goolwa Channel was dominated by agricultural weeds.

Despite the dominance of terrestrial taxa; submergent, amphibious and emergent taxa were present in some wetlands, particularly those that receive runoff from winter rainfall. Goolwa Channel Drive was dominated by emergent and amphibious taxa from 2008 to 2010, particularly at the lowest elevation (the changes in the plant community were generally changes in the abundances of winter annuals at the high elevations). Submergent r-selected species (*Ruppia tuberosa* and *Lamprothamnium macropogon*) were present at the lowest elevation in Loveday Bay Wetland in spring in both years. Submergent k-selected (*Myriophyllum salsugineum*), amphibious *Triglochin sriatum*, *Cyperus gymnocaulos*, *Juncus kraussii*, *Juncus usitatus*) submergent r-selected (*Ruppia polycarpa*) and emergent (*Bolboschoenus caldwellii*, *Eleocharis acuta*, *Phragmites australis*, *Triglochin procerum*) taxa were also present in Milang Wetland in an area that receives stormwater runoff.

Furthermore, wetlands that received rainfall runoff had higher species richness of wetland species and these sites may be important for the persistence of wetland taxa whilst water levels are low.

Despite the current poor condition of lower lakes wetlands there is evidence that there is potential for wetland vegetation to recover if water levels are reinstated to historical levels. Nicol and Ward (2010b) reported that 11 submergent and emergent native taxa were present in the seed banks of Dunn's and Shadow's Lagoons. In addition, the Raukkun Natural Resources Management Team undertook submergent plant seed bank assessments of Narrung, Loveday Bay, Waltowa and Teringie wetlands that showed that all of the aforementioned wetlands had a sediment seed bank (Narrung had the largest and most diverse seed bank). Results from this trial led to Narrung Wetland being watered in spring 2009 and when surveyed in autumn 2010 (whilst the wetland was dry when surveyed) there was evidence that *Ruppia* spp. and charophytes had colonised the wetland and replenished the propagule bank. Monitoring undertaken by Paton and Bailey (2010) in Narrung Wetland confirmed these observations with at least seven species of submergent plants (*Ruppia tuberosa*, *Ruppia polycarpa*, *Lepilaena cylindrocapa*, *Lepilaena preissii*, *Potamogeton pectinatus*, *Lamprothamnium* sp. and *Nitella* sp.) recruiting in response to watering. Furthermore, the plant community that recruited in response to watering was similar to the community described by Holt *et al.* (2005) in spring 2004 (Paton and Bailey 2010). Finally, there was a significant increase in propagule density after the wetland was watered compared to pre-watering densities (Paton and Bailey 2010).

Water levels are currently increasing downstream of Lock 1 due to floods in the northern Murray-Darling Basin but the maximum water level is unclear; therefore, it is unknown which wetlands will reconnect with Lake Alexandrina. However, water levels are expected to decrease over summer; hence, reconnection may only be temporary and wetlands may dry. Reflooding of wetlands will result in mortality of terrestrial taxa (Brock and Casanova 1997; Nicol 2004) and provide suitable conditions for germination of submergent species (Nicol and Ward 2010a; Nicol and Ward 2010b). Nevertheless, if wetlands dry before plants have the opportunity to complete their life cycle the seed bank may become depleted, which may compromise long-term recovery.

4.2 Lakeshores

Sustained low water levels have also resulted in large areas of lake bed, which were historically inundated, becoming exposed. In contrast to wetlands, historical (pre 2007) information regarding the vegetation of lakeshores is not available (except for DEH Biological Survey data (Stewart *et al.* 2009), which generally does not include aquatic habitats). Therefore, comparisons between current and historical conditions cannot be made; however, there is anecdotal evidence that suggests that lakeshores were dominated by emergent taxa and submergent species were

generally absent, except between Clayton and Goolwa and the lower Finniss River and lower Currency Creek (J. Nicol pers. obs.). Similar to wetlands, terrestrial taxa have colonised large areas of lake bed that were historically restricted to elevations above +0.8 m AHD; nevertheless extensive reed beds (generally *Phragmites australis* monocultures) are still present between +0.8 and +0.4 m AHD (Marsland and Nicol 2009).

Similar to 2008-09 the changes in floristic composition through time in Lakes Alexandrina and Albert were generally due changes in abundance of winter annuals but terrestrial taxa have colonised lower elevations than in 2008-09 (Marsland and Nicol 2009). Construction of the Clayton regulator and pumping from Lake Alexandrina has resulted in higher water levels in Goolwa Channel, the lower Finniss River and lower Currency Creek compared with Lakes Alexandrina and Albert. Higher water levels in Goolwa Channel have resulted in different plant communities developing between -0.5 m and +0.8 m AHD in Goolwa Channel compared with Lakes Alexandrina and Albert in 2009-10.

In 2009-10 in Goolwa Channel, the lower Finniss River and lower Currency Creek zonation of the plant community in relation to water depth has occurred (*sensu* Spence 1982), whereas in 2008-09 the fringing reed beds were present between +0.4 and +0.8 m AHD but the lower elevations were dominated by terrestrial taxa and bare soil (Marsland and Nicol 2009). In 2009-10 at high elevations (+0.4 to +0.8 m AHD) in Goolwa Channel the plant community was dominated by emergent and amphibious species such as *Phragmites australis*, *Muehlenbeckia florulenta*, *Typha domingensis* and *Calystegia sepium*. The corresponding elevations in Lakes Alexandrina and Albert were dominated by *Phragmites australis* and terrestrial taxa. At intermediate elevations (0 to +0.4 m AHD) in 2009-10 emergent species that are adapted to deeper water such as *Typha domingensis* and *Schoenoplectus validus* were common in Goolwa Channel but in Lakes Alexandrina and Albert terrestrial species were the dominant functional group.

At -0.5 m AHD in 2009-10 only submergent and amphibious species (*Potamogeton pectinatus* and *Myriophyllum salsugineum*) were present in Goolwa Channel. In addition, *Vallisneria australis*, *Ruppia megacarpa*, *Ceratophyllum demersum* and *Ruppia polycarpa* were also observed in inundated areas of Goolwa Channel but not present in any quadrats. Submergent species were not observed until autumn 2010, which was not unexpected because the spring 2009 survey was undertaken four weeks after pumping ceased and the majority of submergent species require longer than four weeks of inundation to germinate (Nicol 2004; Nicol and Ward 2010a; Nicol and Ward 2010b). The colonisation of submergent species in Goolwa Channel in response to regulated inundation provides further evidence that the system is resilient and the aquatic plant community has the capacity to recover from low water levels. All submergent species observed in Goolwa Channel (except *Ceratophyllum demersum*) were present in the sediment seed bank (Nicol and Ward 2010b),

which suggests that the seed bank is an important source of propagules for recolonisation of submergent species. In Lakes Alexandrina and Albert bare soil and terrestrial species were dominant at -0.5 m AHD.

Furthermore, the majority of the terrestrial taxa that have colonised the dry lakebed in Lakes Alexandrina and Albert were exotics and there was also an increase in the number of exotics compared to 2008-09 (Marsland and Nicol 2009). The dominance of exotic species is probably due to adjacent land use (agricultural and urban), which is dominated by exotic species.

Surface water salinity in Goolwa Channel (which was in excess of 19,000 $\mu\text{S}\cdot\text{cm}^{-1}$ adjacent to the Hindmarsh Island Bridge in March 2010) was higher than the reported tolerances for several of the emergent and submergent species present (Bailey *et al.* 2002). Extensive stands of healthy *Typha domingensis* (maximum reported salinity tolerance of 8,000 $\mu\text{S}\cdot\text{cm}^{-1}$), *Phragmites australis* (reported to show signs of severe stress at 15,000 $\mu\text{S}\cdot\text{cm}^{-1}$) and *Schoenoplectus validus* (maximum reported salinity tolerance of 700 $\mu\text{S}\cdot\text{cm}^{-1}$) (Bailey *et al.* 2002) were present in Goolwa Channel despite the elevated salinity. These results provide evidence that there are local salt tolerant ecotypes of the aforementioned species present in Goolwa Channel and potentially more broadly in the Lower Lakes.

Similar to wetlands, the impact of the current rise in water levels on lakeshores will depend on the maximum height and duration of inundation, and will only affect Lake Alexandrina (although water levels in Lake Albert may rise due to pumping). Submergent species may recruit in sheltered areas where these species were present historically (i.e. near Milang Jetty J. Nicol pers. obs.); however, hydroperiods need to be sufficient for species to complete their life cycle or the seed bank may become depleted. Results from Goolwa Channel showed that diverse fringing communities developed in response to high water levels in areas that were almost *Phragmites australis* monocultures (Marsland and Nicol 2009; Gehrig and Nicol 2010). The rise in water level may provide conditions suitable for regeneration of fringing emergent and amphibious species and increase diversity of the fringing plant community.

4.3 TLM Targets

Continued low water levels have meant that TLM target V3 has not been met for Lakes Alexandrina or Albert (except Narrung Wetland) due to the dominance of terrestrial species. However, the construction of the Clayton regulator and resultant high water levels has meant that aquatic and floodplain vegetation has recolonised areas between Goolwa and Clayton and the lower Finniss River and Currency Creek. Goolwa Channel monitoring (Gehrig and Nicol 2010), seed bank assessment (Nicol and Ward 2010a; Nicol and Ward 2010b) results and observations at Narrung Wetland have shown that there is capacity for wetland vegetation to

recolonise areas that have been dry for up to 3 years. Increased water levels in spring 2010 may also result in recolonisation of aquatic, amphibious and emergent taxa; however, the maximum water level and in turn the potential for wetland plant recolonisation is unclear at this stage. If water is available, the potential changes in the fringing plant community in Lake Alexandrina may occur in Lake Albert if water levels can be maintained at a similar level by pumping.

4.4 Further Studies

- TLM and Goolwa Channel vegetation monitoring for 2010-11 have been funded, which will provide information regarding the impact of increased water levels in wetlands and lakeshores. In addition, this monitoring will provide information on succession of plant communities in Goolwa Channel under managed inundation.
- An honours research project investigating the impacts of acid sulfate soil and remediation measures (e.g. liming and mulching) on the performance of emergent species will be undertaken by Rod Ward and supervised by Molly Whelan, Duncan McKay and Jason Nicol at Flinders University in 2010-11.
- Results from targeted seed bank studies, Goolwa Channel monitoring and observation in Narrung Wetland have shown that there is the capacity for submergent plant communities to recover if water levels are restored to historical levels. Nevertheless, the spatial extent of these studies is limited and the capacity of vegetation to recover in other areas of the Lower Lakes is unknown. In addition, there is no information regarding seed longevity under different physico-chemical conditions (e.g. salinity and pH).
- The seed bank is only one mechanism of recovery of plant communities, there is no information regarding other mechanisms of recovery such as dispersal of plant propagules into wetlands by water (hydrochory) or animals (zoochory).
- Surface water salinity in Goolwa Channel exceeded 19,000 $\mu\text{S}\cdot\text{cm}^{-1}$ in March 2010, which is higher than the reported tolerances of many of the species that were present and appeared to be in very good condition (Bailey *et al.* 2002). Therefore, published salinity tolerances probably do not apply to many species in the Lower Lakes region and the tolerances of local ecotypes are unknown.

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6 Appendices

Appendix 1: GPS coordinates (UTM format, map datum WGS84) for lakeshore and wetland understorey vegetation monitoring sites.

| Site | Easting | Northing | Site type |
|-------------------------------|---------|----------|-----------|
| Bremer Mouth Lakeshore | 323061 | 6081991 | lakeshore |
| Brown Beach 1 | 350172 | 6052777 | lakeshore |
| Brown Beach 2 | 350287 | 6053158 | lakeshore |
| Clayton Bay | 311301 | 6070626 | lakeshore |
| Currency Creek 3 | 296772 | 6074222 | lakeshore |
| Currency Creek 4 | 301013 | 6071800 | lakeshore |
| Goolwa North | 303330 | 6070156 | lakeshore |
| Goolwa South | 300490 | 6066366 | lakeshore |
| Hindmarsh Island Bridge 01 | 299670 | 6068521 | lakeshore |
| Hindmarsh Island Bridge 02 | 299695 | 6068616 | lakeshore |
| Lake Reserve Rd | 339298 | 6089987 | lakeshore |
| Loveday Bay | 329431 | 6058407 | lakeshore |
| Loveday Bay Lakeshore | 326621 | 6061647 | lakeshore |
| Lower Finniss 02 | 305131 | 6076401 | lakeshore |
| Milang | 315964 | 6079870 | lakeshore |
| Milang Lakeshore | 316081 | 6079746 | lakeshore |
| Pt Sturt Lakeshore | 322811 | 6069643 | lakeshore |
| Pt Sturt Water Reserve | 317673 | 6070784 | lakeshore |
| Teringie Lakeshore | 327461 | 6066887 | lakeshore |
| Upstream of Clayton Regulator | 312281 | 6069151 | lakeshore |
| Wally's Landing | 303066 | 6079631 | lakeshore |
| Warrenjie 1 | 347722 | 6049163 | lakeshore |
| Lower Finniss 03 | 305131 | 6072406 | lakeshore |
| Narrung Lakeshore | 333762 | 6069807 | lakeshore |
| Nurra Nurra | 341786 | 6063837 | lakeshore |
| Warrenjie 2 | 348487 | 6049133 | lakeshore |
| Angas Mouth | 318391 | 6081206 | wetland |
| Bremer Mouth | 323056 | 6082019 | wetland |
| Dunns Lagoon | 312417 | 6070300 | wetland |
| Goolwa Channel Drive | 307024 | 6064437 | wetland |
| Hunters Creek | 308219 | 6065526 | wetland |
| Poltalloch | 343248 | 6071554 | wetland |
| Pt Sturt | 322778 | 6069794 | wetland |
| Teringie | 327334 | 6065286 | wetland |
| Waltowa | 353908 | 6057756 | wetland |
| Narrung | 334542 | 6068744 | wetland |

Appendix 2: Species list, functional classification (Gehrig and Nicol 2010), life history strategy and conservation status (state conservation status from listings in Barker *et al.* (2005) and regional conservation status from listings in Lang and Kraehenuehl (2001)) from all sites and survey dates (*denotes exotic taxon).

| Taxa | Functional Group | Life history strategy | Status and Comments |
|----------------------------------|---|-----------------------|--|
| <i>Acacia myrtifolia</i> | Terrestrial dry | Perennial | Native |
| <i>Agapanthus praecox</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Apium graveolens</i> * | Terrestrial damp | Annual | Exotic |
| <i>Arctotheca calendula</i> * | Terrestrial dry | Annual | Exotic |
| <i>Asparagus asparagoides</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Asparagus officinalis</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Aster subulatus</i> * | Terrestrial damp | Annual | Exotic |
| <i>Atriplex prostrata</i> * | Terrestrial damp | Perennial | Exotic |
| <i>Atriplex semibaccata</i> | Terrestrial dry | Perennial | Native-Listed as Uncommon in the Murray Region |
| <i>Atriplex suberecta</i> | Floodplain | Perennial | Native |
| <i>Avena</i> spp.* | Terrestrial dry | Annual | Exotic- <i>Avena</i> spp. is comprised of <i>Avena barbata</i> and <i>Avena fatua</i> |
| <i>Berula erecta</i> | Emergent | Perennial | Native |
| <i>Bolboschoenus caldwellii</i> | Emergent | Perennial | Native |
| <i>Brassica rapa</i> * | Terrestrial dry | Annual | Exotic |
| <i>Brassica tournifortii</i> * | Terrestrial dry | Annual | Exotic |
| <i>Briza minor</i> * | Terrestrial dry | Annual | Exotic |
| <i>Bromus diandrus</i> * | Terrestrial dry | Annual | Exotic |
| <i>Bromus hordeaceus</i> * | Terrestrial dry | Annual | Exotic |
| <i>Bromus unioloides</i> * | Terrestrial dry | Annual | Exotic |
| <i>Calystegia sepium</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native-Listed as Uncommon in the Murray and Southern Lofty Regions |
| <i>Carpobrotus rossii</i> | Terrestrial dry | Perennial | Native |
| <i>Centaureum tenuiflorum</i> * | Terrestrial damp | Annual | Exotic |
| <i>Centaurea calcitrapa</i> * | Terrestrial damp | Annual | Exotic |
| <i>Chenopodium album</i> * | Terrestrial damp | Annual | Exotic |
| <i>Chenopodium glaucum</i> * | Terrestrial damp | Annual | Exotic |
| <i>Chenopodium nitrariaceum</i> | Terrestrial dry | Perennial | Native |
| <i>Conyza bonariensis</i> * | Terrestrial damp | Annual | Exotic |
| <i>Cotula coronopifolia</i> * | Amphibious fluctuation responder-plastic | Perennial | Exotic |
| <i>Crinum</i> sp.* | Terrestrial dry | Perennial | Exotic-garden escapee not in any of the identification keys and could not be identified to species |
| <i>Cyperus exaltatus</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Cyperus gymnocaulos</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Distichlis distichophylla</i> | Terrestrial damp | Perennial | Native-Listed as Uncommon in the Murray Region |
| <i>Disphyma crassifolium</i> | Terrestrial dry | Perennial | Native |
| <i>Ehrharta longiflora</i> * | Terrestrial damp | Annual | Exotic |
| <i>Einadia nutans</i> | Terrestrial dry | Perennial | Native |
| <i>Eleocharis acuta</i> | Emergent | Perennial | Native |
| <i>Enchylaena tomentosa</i> | Terrestrial dry | Perennial | Native |
| <i>Epilobium pallidiflorum</i> | Terrestrial damp | Perennial | Native-Listed as Uncertain in the Murray Region and uncommon in the Southern Lofty Region |
| <i>Eragrostis australasica</i> | Floodplain | Perennial | Native |

| Taxa | Functional Group | Life history strategy | Status and Comments |
|----------------------------------|--|-----------------------|--|
| <i>Eragrostis curvula</i> * | Terrestrial damp | Annual | Exotic-Proclaimed pest plant in SA |
| <i>Eragrostis</i> sp. | Terrestrial damp | Annual | Native-could not identify to species |
| <i>Euphorbia terracina</i> * | Terrestrial dry | Annual | Exotic-Proclaimed pest plant in SA |
| <i>Foeniculum vulgare</i> * | Terrestrial damp | Annual | Exotic |
| <i>Frankenia pauciflora</i> | Terrestrial dry | Perennial | Native |
| <i>Fumaria bastardii</i> * | Terrestrial damp | Annual | Exotic |
| <i>Gahnia filum</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native-Listed as Rare in the Murray and Southern Lofty Regions |
| <i>Glyceria australis</i> | Emergent | Perennial | Native |
| <i>Halosarcia pergranulata</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Heliotropium europaeum</i> * | Floodplain | Annual | Exotic |
| <i>Holcus lanatus</i> * | Terrestrial damp | Annual | Exotic |
| <i>Hordeum vulgare</i> * | Terrestrial dry | Annual | Exotic |
| <i>Hydrocotyle verticillata</i> | Amphibious fluctuation responder-plastic | Perennial | Native-Listed as Uncertain in the Southern Lofty Region |
| <i>Hypochoeris glabra</i> * | Terrestrial dry | Annual | Exotic |
| <i>Hypochoeris radicata</i> * | Terrestrial dry | Annual | Exotic |
| <i>Isolepis nodosa</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Isolepis platycarpa</i> | Amphibious fluctuation tolerator-low growing | Perennial | Native |
| <i>Isolepis</i> sp. | Amphibious fluctuation tolerator-low growing | Perennial | Native-could not identify to species |
| <i>Juncus acutus</i> * | Amphibious fluctuation tolerator-emergent | Perennial | Exotic |
| <i>Juncus kraussii</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Juncus usitatus</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Lachnagrostis filiformis</i> | Floodplain | Annual | Native |
| <i>Lactuca saligna</i> * | Terrestrial dry | Annual | Exotic |
| <i>Lactuca serriola</i> * | Terrestrial dry | Annual | Exotic |
| <i>Lagurus ovatus</i> * | Terrestrial dry | Annual | Exotic |
| <i>Lamprothamnium macropogon</i> | Submergent r-selected | Annual | Native |
| <i>Lobelia alata</i> | Terrestrial damp | Perennial | Native |
| <i>Lolium</i> spp.* | Terrestrial dry | Annual | Exotic- <i>Lolium</i> spp. comprises of <i>Lolium perenne</i> and <i>Lolium rigidum</i> |
| <i>Lupinus cosentinii</i> * | Terrestrial dry | Annual | Exotic |
| <i>Lycium ferocissimum</i> * | Terrestrial dry | Perennial | Exotic-Proclaimed pest plant in SA |
| <i>Lycopus australis</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native-Listed as Rare in the Murray Region |
| <i>Malva parviflora</i> * | Terrestrial dry | Annual | Exotic |
| <i>Medicago</i> spp.* | Terrestrial dry | Annual | Exotic- <i>Medicago</i> spp. comprises of <i>Medicago polymorpha</i> , <i>Medicago truncatula</i> and <i>Medicago minima</i> |
| <i>Melaleuca halmaturorum</i> | Amphibious fluctuation tolerator-woody | Perennial | Native |
| <i>Melilotus indica</i> * | Terrestrial dry | Annual | Exotic |
| <i>Mentha australis</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Mentha</i> spp.* | Amphibious fluctuation tolerator-emergent | Perennial | Exotic- <i>Mentha</i> spp. comprises of <i>Mentha piperita</i> , <i>Mentha pulegium</i> and <i>Mentha spicata</i> |
| <i>Mimulus repens</i> | Amphibious fluctuation tolerator-low growing | Perennial | Native |
| <i>Muehlenbeckia florulenta</i> | Amphibious fluctuation tolerator-woody | Perennial | Native |
| <i>Muehlenbeckia gunnii</i> | Amphibious fluctuation tolerator-woody | Perennial | Native |
| <i>Myriophyllum salsugineum</i> | Submergent k-selected | Perennial | Native-Listed as Uncertain in the Southern Lofty Region |

| Taxa | Functional Group | Life history strategy | Status and Comments |
|------------------------------------|--|-----------------------|--|
| <i>Onopordum acanthium</i> * | Terrestrial damp | Annual | Exotic |
| <i>Oxalis pes-caprae</i> * | Terrestrial dry | Annual | Exotic-Proclaimed pest plant in SA |
| <i>Paspalum distichum</i> * | Terrestrial damp | Perennial | Exotic |
| <i>Pennisetum clandestinum</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Persicaria lapathifolia</i> | Amphibious fluctuation responder-plastic | Perennial | Native |
| <i>Phalaris arundinacea</i> * | Amphibious fluctuation tolerator-emergent | Perennial | Exotic |
| <i>Phragmites australis</i> | Emergent | Perennial | Native |
| <i>Picris hieracoides</i> | Terrestrial dry | Annual | Native |
| <i>Plantago coronopus</i> * | Terrestrial dry | Annual | Exotic |
| <i>Plantago lanceolata</i> * | Terrestrial dry | Annual | Exotic |
| <i>Plantago major</i> * | Terrestrial dry | Annual | Exotic |
| <i>Polygogon monspeliensis</i> * | Amphibious fluctuation tolerator-emergent | Annual | Exotic |
| <i>Polygonum aviculare</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Potamogeton pectinatus</i> | Submergent k-selected | Perennial | Native |
| <i>Pseudognaphalium luteoalbum</i> | Floodplain | Annual | Native |
| <i>Puccinellia</i> sp.* | Terrestrial damp | Annual | Exotic-could not be identified to species but was not <i>Puccinellia stricta</i> or <i>Puccinellia perluxa</i> |
| <i>Reichardia tingitana</i> * | Terrestrial dry | Annual | Exotic |
| <i>Rhagodia spinescens</i> | Terrestrial dry | Perennial | Native |
| <i>Ruppia polycarpa</i> | Submergent r-selected | Annual | Native |
| <i>Ruppia tuberosa</i> | Submergent r-selected | Annual | Native |
| <i>Salix babylonica</i> * | Emergent | Perennial | Exotic |
| <i>Salsola kali</i> | Terrestrial dry | Perennial | Native |
| <i>Samolus repens</i> | Terrestrial damp | Perennial | Native- Listed as Rare in the Murray Region and Uncommon the Southern Lofty Region |
| <i>Sarcocornia quinqueloflora</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Scabiosa atropurpurea</i> * | Terrestrial dry | Annual | Exotic |
| <i>Schoenoplectus pungens</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native-Listed as Rare in the Southern Lofty Region |
| <i>Schoenoplectus validus</i> | Emergent | Perennial | Native |
| <i>Sclerolaena blackiana</i> | Terrestrial dry | Perennial | Native-Listed as Rare in SA |
| <i>Senecio cunninghamii</i> | Floodplain | Perennial | Native |
| <i>Senecio pterophorus</i> * | Terrestrial dry | Annual | Exotic |
| <i>Senecio runcinifolius</i> | Floodplain | Perennial | Native-Listed as Uncommon in the Murray Region |
| <i>Solanum nigrum</i> * | Terrestrial damp | Annual | Exotic |
| <i>Sonchus asper</i> * | Terrestrial damp | Annual | Exotic |
| <i>Sonchus oleraceus</i> * | Terrestrial damp | Annual | Exotic |
| <i>Spergularia marina</i> * | Terrestrial damp | Annual | Exotic |
| <i>Suaeda australis</i> | Amphibious fluctuation tolerator-emergent | Perennial | Native |
| <i>Silybum marianum</i> * | Terrestrial damp | Annual | Exotic-Proclaimed pest plant in SA |
| <i>Tamarix aphylla</i> * | Terrestrial dry | Perennial | Exotic |
| <i>Trifolium</i> spp.* | Terrestrial dry | Annual | Exotic- <i>Trifolium</i> spp. comprises of <i>Trifolium angustifolium</i> , <i>Trifolium arvense</i> , <i>Trifolium repens</i> and <i>Trifolium subterraneum</i> |
| <i>Triglochin procerum</i> | Emergent | Perennial | Native-Listed as Uncommon in the Southern Lofty Region |
| <i>Triglochin striatum</i> | Amphibious fluctuation tolerator-low growing | Perennial | Native |
| <i>Triticum</i> sp.* | Terrestrial dry | Annual | Exotic-could not be identified to species |

| Taxa | Functional Group | Life history strategy | Status and Comments |
|------------------------------|-------------------------|------------------------------|--|
| <i>Typha domingensis</i> | Emergent | Perennial | Native |
| <i>Urtica urens</i> * | Terrestrial damp | Annual | Exotic |
| <i>Vallisneria australis</i> | Submergent k-selected | Perennial | Native-Listed as Uncommon in the Murray Region and Threatened in the Southern Lofty Region |
| <i>Vicia sativa</i> * | Terrestrial dry | Annual | Exotic |
| <i>Wilsonia rotundifolia</i> | Terrestrial damp | Perennial | Native |

Appendix 3: Species present in wetlands from spring 2008 to autumn 2010 (*denotes exotic species).

| Species | Angas Mouth | Bremer Mouth | Dunn's Lagoon | Goolwa Channel Drive | Hunter's Creek | Loveday Bay | Milang | Narrung | Poltalloch | Point Sturt | Teringie | Waitowa |
|----------------------------------|-------------|--------------|---------------|----------------------|----------------|-------------|--------|---------|------------|-------------|----------|---------|
| <i>Agapanthus praecox</i> * | | | | | | | * | | | | | |
| <i>Asparagus asparagoides</i> * | * | | | | | | * | | | | | |
| <i>Aster subulatus</i> * | | * | * | | | * | * | | | * | | |
| <i>Atriplex prostrata</i> * | * | | * | * | * | * | * | | | * | | |
| <i>Atriplex semibaccata</i> | | | | | | | | * | | * | | |
| <i>Atriplex suberecta</i> | | | | | | | | | | * | | |
| <i>Avena</i> spp.* | | | * | | | | * | | | * | | * |
| <i>Berula erecta</i> | | * | | | | | | | | | | |
| <i>Bolboschoenus caldwellii</i> | | | | * | * | * | * | | * | | | |
| <i>Brassica rapa</i> * | | | * | | | | | | | | | |
| <i>Brassica tournefortii</i> * | | | | | | | | | | | * | |
| <i>Bromus diandrus</i> * | | | * | | | | * | | | * | * | * |
| <i>Bromus hordeaceus</i> * | | | * | | | | * | | | * | * | * |
| <i>Bromus unioloides</i> * | | | | | | | | | | | | * |
| <i>Calystegia sepium</i> | | | * | | | | | | | | | |
| <i>Carpobrotus rossii</i> | | | | | | | | | | | | * |
| <i>Centaurea calcitrapa</i> * | | * | | * | | * | * | | | | | |
| <i>Chenopodium album</i> * | * | | * | | | | * | | | | | |
| <i>Chenopodium glaucum</i> * | | | * | | | | | | * | * | | |
| <i>Conyza bonariensis</i> * | | | * | * | | | | | | | | |
| <i>Cotula coronopifolia</i> * | | * | * | * | | * | * | | * | * | * | * |
| <i>Crinum</i> sp.* | | | | | | | * | | | | | |
| <i>Cyperus exaltatus</i> | * | | | | | | | | | | | |
| <i>Cyperus gymnocaulos</i> | | | * | | | * | * | | * | * | * | |
| <i>Disphyma crassifolium</i> | | | | | | | | | | | | * |
| <i>Distichlis distichophylla</i> | | | * | * | * | * | * | * | * | * | * | |
| <i>Ehrharta longiflora</i> * | | | | * | | | * | | | | | |
| <i>Einadia nutans</i> | | | | | | | | | | * | | |
| <i>Eleocharis acuta</i> | | | | | | | * | | | | | |
| <i>Eleocharis tomentosa</i> | | | * | | | | * | | | * | * | |
| <i>Eragrostis curvula</i> * | | | * | | | * | * | * | * | * | * | * |
| <i>Eragrostis</i> sp. | | | * | | | | * | * | * | * | * | * |
| <i>Frankenia pauciflora</i> | | | | * | | * | * | * | | * | * | * |

| Species | Angas Mouth | Bremer Mouth | Dunn's Lagoon | Goolwa Channel Drive | Hunter's Creek | Loveday Bay | Milang | Narrung | Potalloch | Point Sturt | Teringie | Waitowa |
|-------------------------------------|--------------|--------------|---------------|----------------------|----------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|
| <i>Phragmites australis</i> | * | * | * | * | | | * | * | | | | * |
| <i>Picris hieracoides</i> | | | * | | | | | | | | | |
| <i>Plantago coronopus*</i> | | * | * | * | | * | * | | | * | * | * |
| <i>Polygonum aviculare*</i> | | | | | | | | | | * | | |
| <i>Polypogon monspeliensis*</i> | * | * | | | | | * | * | * | * | * | |
| <i>Psuedognaphalium luteo-album</i> | | | * | | | | | * | | | | |
| <i>Puccinellia sp.*</i> | | | | | | | | | | | | |
| <i>Reichardia tingitana*</i> | | * | * | * | | | * | | | * | | * |
| <i>Ruppia polycarpa</i> | | | | | | | * | | | | | |
| <i>Ruppia tuberosa</i> | | | | | * | | | | | | | |
| <i>Salsola kali</i> | | | | | | | | | | | * | |
| <i>Samolus repens</i> | | | * | * | | * | * | | | * | * | |
| <i>Sarcocornia quinqueflora</i> | | | * | * | * | * | * | * | * | * | * | * |
| <i>Schoenoplectus pungens</i> | | | * | * | * | * | * | | | | | |
| <i>Senecio cunninghamii</i> | | * | | | | | * | * | | | * | * |
| <i>Senecio pterophorus*</i> | | * | * | | | * | * | * | | * | | |
| <i>Senecio runcinifolius</i> | | | * | | | | | | | | | * |
| <i>Sonchus asper*</i> | | | * | | | | | | | | | |
| <i>Sonchus oleraceus*</i> | * | * | * | * | * | * | * | * | * | * | * | * |
| <i>Spergularia marina*</i> | | | * | * | | | * | * | * | * | * | * |
| <i>Suaeda australis</i> | | | * | * | * | * | * | * | * | * | * | * |
| <i>Trifolium spp.*</i> | | * | * | | | * | * | | | * | * | |
| <i>Triglochin procerum</i> | | * | * | | | | | | | | | |
| <i>Triglochin striatum</i> | * | * | * | | | | * | | | | * | |
| <i>Typha domingensis</i> | * | * | * | * | | * | * | | | | | |
| <i>Urtica urens*</i> | | | * | | | | | | | | | |
| <i>Vallisneria australis</i> | | * | | | | | | | | | | |
| <i>Vicia sativa*</i> | | | | | | | * | | | | | |
| <i>Wilsonia rotundifolia</i> | | | * | | | * | * | | | | | |
| Species Richness | 12 | 21 | 57 | 22 | 11 | 30 | 53 | 18 | 18 | 34 | 30 | 23 |
| No. Exotics | 7 | 13 | 30 | 10 | 3 | 16 | 33 | 9 | 11 | 23 | 18 | 14 |
| % Exotics | 58.33 | 61.90 | 52.63 | 45.45 | 27.27 | 53.33 | 62.26 | 50.00 | 61.11 | 67.65 | 60.00 | 60.87 |

